Waste Plastics' Liquefaction into Fuel Fraction

Subjects: Polymer Science

Contributor: Anna Matuszewska , Marlena Owczuk , Krzysztof Biernat

Polymers and plastics are crucial materials in many sectors of our economy, due to their numerous advantages. They also have some disadvantages, among the most important are problems with the recycling and disposal of used plastics. The recovery of waste plastics is increasing every year, but over 27% of plastics are landfilled. The rest is recycled, where, unfortunately, incineration is still the most common management method. From an economic perspective, waste management methods that lead to added-value products are most preferred—as in the case of material and chemical recycling. Since chemical recycling can be used for difficult wastes (poorly selected, contaminated), it seems to be the most effective way of managing these materials. Moreover, as a result this of kind of recycling, it is possible to obtain commercially valuable products, such as fractions for fuel composition and monomers for the reproduction of polymers.

plastic waste

liquefaction

fuels

quality standards

1. Introduction

Access to cheap energy carriers is crucial for the development of many sectors of the economy worldwide. The constantly growing demand for these carriers contributes to the depletion of the world's oil, gas and coal resources. This growing demand is observed in both developed and developing countries. According to the International Energy Agency report ^[1], the world's total consumption of crude oil was at 150.1, 169.3 and 153.5 exajoules, while natural gas consumption was 57.2, 69.7 and 68.2 exajoules, in 2010, 2019 and 2020, respectively. The recorded reduction in fuel consumption in 2020 was caused by the global COVID-19 pandemic. The increasing demand for energy carriers will entail an increase in the prices of these goods. According to the Report of U.S. Energy Information Administration (EIA) ^[2], the price of oil in November 2021 was 81 USD per barrel, while in the corresponding period of 2020 it was 43 USD. Over the course of 2021, oil prices increased due to the use of global oil reserves at an average of 1.4 million barrels per day. The successive coronavirus waves during the global pandemic situation have contributed greatly to the imbalance in the use and price of oil resources ^{[3][4]}. Despite the slowdown in global oil consumption due to concerns about other COVID-19 variants, the EIA predicts that prices per barrel will remain at a similar level as in 2021.

The growing demand for and rising prices of energy have led to a crisis in the fuel market and the search for substitute fossil fuel carriers. The second reason for replacing fossil fuels with raw materials from renewable sources is the global trend of decarbonizing economies ^[5]. The aim of decarbonization is to counteract severe climate change, including global warming ^[6]; as recommended by the Intergovernmental Panel on Climate Change,

to avoid the aggravation of the negative effects of climate change, a temperature rise limit of 1.5 °C should be achieved [Z].

One way of minimizing the effects of the energy and climate crisis is to, inter alia, move away from a linear economy (a vision called Cradle-to-Grave (CtG)), in which the last stage of a product's life is waste, in favor of a circular economy (CE). The CE treats waste as a valuable raw material and is based on the Cradle-to-Cradle (CtC) model. It is a global strategy for sustainable development, which is assumed to be renovating or regenerative in nature ^[8]. It replaces the 'end of life' concept with the renovation/regeneration of the use process and moves towards the use of renewable energy. It aims to minimize the amount of generated waste, and products, materials and raw materials should remain in the economic cycle for as long as possible ^[9].

2. Plastic Waste Management

The problem with the management of plastics waste is already global, but the situation is improving every year. This is evidenced, inter alia, by the annual data published in the report of the Plastics Europe Foundation, which indicates that, in the European Union in 2020, compared to 2006, recycling of plastic waste increased by almost 118%, energy recovery through incineration increased by 77% and the amount of waste sent to landfills decreased by over 46% ^[10]. In 2020, over 29 million tonnes of post-consumer plastic waste were collected in European Union countries. Unfortunately, as much as 23.4% of them went to landfills, and only 34.6% was subjected to typical recycling. A total of 42% of collected waste was utilized through energy recovery through incineration (energy recycling) ^[10]. This form of recycling (in accordance with the principle of waste management hierarchy ^[11]) should only be used as a last resort for waste that is not suitable for other forms of recycling, such as depolymerization, material or chemical recycling. Sadly, the above data show that waste incineration is the most common and dominant form of PW management, which unfortunately leads to the complete elimination of plastic material from the economy and makes it impossible to use this material in any other way. In addition, large amounts of harmful and toxic compounds are generated during incineration ^[12].

The still relatively low level of PW recycling (depolymerization, material and chemical recycling) results from the complexity of the processing methods, significant energy demand or/and unsatisfactory quality of the obtained products. The best method for managing plastics is their depolymerization to monomers (input ingredients for the production of plastics). However, this method is applicable, with good efficiency, to a few, selected polymers (polyethylene terephthalate, polymethyl methacrylate, polystyrene) ^{[13][14][15]}. In addition, the raw material must be clean and free from any additives of other materials. Unfortunately, this method cannot be applied to polyolefins, which accounts for approximately 70–76% of the waste plastic stream ^[16].

In the case of polyolefins, the best method for their management is material (mechanical) recycling to regranulate. This method is considered the most environmentally friendly due to the low value of CO_2 emissions compared to other methods ^[17], but it is not free from drawbacks. The feedstock for processing must be very well selected and clean, and unfortunately the material loses quality with each subsequent processing cycle. In addition, the use of different types of plastics on various parts of packaging (e.g., labels, caps, packaging) and the use of additives and

fillers for plastics make it difficult to recycle and recover individual groups of materials and reduce the quality of the obtained final products ^{[18][19]}. Post-industrial plastics (such as cuttings and trimmings from production, runners from injection moulding or granulation residues) are best-suited for such recycling because they have a known and homogeneous composition and contain few impurities ^{[20][21]}.

Regarding post-consumer PW, which is a complex mixture of different plastics and other materials, a sorting process should be carried out to obtain separate components should. Selected components, depending on their quality, can be mechanically recycled (for example, into outdoor furniture, flower pots, elements of terraces), directed to chemical recycling or sent to incineration ^[22]. However, due to the difficulty of completely separating the individual materials and contaminations of post-consumer waste, chemical recycling methods seem to be the most appropriate for their management ^[23]. **Figure 1** shows a schematic diagram of the main directions of plastic waste management.



Figure 1. The plastic waste management recycling technologies.

Plastic waste is a valuable raw material for recycling, but the converting method is influenced by the economic efficiency of the process, which depends, among other factors, on the material composition, the size and quality (homogeneity, purity) of the collected waste stream, the availability of sorting and processing technologies, and the market demand for the products obtained as a result of recycling.

Due to the significant share of polyolefins and polystyrene in the post-consumer waste stream, they are a valuable material for the production of engine fuels ^[24]. **Figure 2** shows the main directions of the thermochemical conversion of polyolefins and polystyrene to hydrocarbons, most frequently encountered in the literature.



Figure 2. The main directions of thermochemical conversion of polyolefins and polystyrene to hydrocarbons.

3. Pyrolysis

One of the methods of converting waste plastics to hydrocarbon fractions is pyrolysis. This enables the transformation of high-molecular compounds, such as waste plastics, into compounds with much shorter carbon chains. It is considered a promising method for the conversion of addition polymers such as polyethylene and polypropylene (polyolefins) ^{[25][26]}. It offers the opportunity to recycle this polyolefin stream, which cannot be economically managed in conventional mechanical recycling, thus avoiding their incineration and the generation of hazardous substances such as dioxins and furans ^{[27][28]}. The pyrolysis oil (a mixture of waxes and a liquid fraction) obtained as a result of pyrolysis can be used as a substitute for heavy fuel oil or undergo further processing, e.g., in the petrochemical industry ^{[29][30]}. Furthermore, as shown in Reference ^[31], large-scale pyrolysis is economically viable. For example, in the case of an investment in installation that processes 100,000 kg/h of waste plastics, income is generated within the first year and has a positive net present value. The advantage of pyrolysis is that pre-segregated, mixed and contaminated plastic waste with both organic and inorganic material can be subjected to this process ^{[32][33]}.

The pyrolysis process is based on the thermochemical decomposition of various materials of organic and synthetic origin at elevated temperatures and in an oxygen-free atmosphere. Typically, the process is carried out at temperatures in the range of 500–800 °C and, in the case of plastics, it follows four types of mechanism: end-chain scission or depolymerization, random-chain scission, chain stripping and cross-linking ^[34]. The main products of the pyrolysis process are pyrolysis oil, gaseous products and the solid fraction (char). On average, depending on the technology used, 45–50% of the oil fraction, 35–40% of the gaseous fraction and 10–20% of char is obtained in the process ^[35]. It was generally found that with the increase of the process temperature, the yield of the liquid fraction increases, while the yield of the solid fraction decreases ^[36]. However, for selected plastics (e.g., for PE over catalyst), it is possible to obtain a liquid fraction in an amount exceeding even 90% ^{[37][38]}. The high propensity of plastics to generate significant amounts of oil fraction during pyrolysis is the result of high volatile matter and low

ash content in relation to other materials (e.g., biomass). These two parameters are the key factors influencing the pyrolysis oil yield ^[39].

Due to the process conditions (vapor residence time in the reactor, temperature, raw material heating rate), pyrolysis is divided into three types [40][41]:

- Slow pyrolysis—process temperature from 350 °C to 550 °C, heating rate from 1 to 10 °C/min, extended steam
 residence time, where char is the main product's first bullet;
- Fast pyrolysis—the process temperature usually ranges from 500 °C to 700 °C, the raw material heating rate is above 100 °C/min, and the vapor residence time is usually within a few seconds; the main product is the liquid fraction, and in the case of polyolefin, pyrolysis also waxes;
- Flash pyrolysis—the process temperature usually exceeds 700 °C, the heating rate of the raw material >200 °C/s, and the vapor residence time is in the millisecond range.

In addition to the process time, pyrolysis can also be divided depending on other conditions of the process (e.g., presence of a catalyst, type of gas) or the type of supporting equipment ^[42].

The composition of pyrolysis oil depends not only on the process conditions, but also on the type of waste plastic in the raw material. A higher yield of aromatic hydrocarbons is obtained when PS and PET are pyrolyzed. On the other hand, PP and PE produces oil that is rich in paraffins, olefins and waxes based on aliphatic hydrocarbons. Wax is an intermediate product, mainly composed of long-chain hydrocarbons (C20+) with a high boiling point. In order to transform it into fuel products, it is necessary to separate it after pyrolysis and subject it to further cracking ^[40]. The pyrolysis oil also requires further processing (e.g., distillation, refining, hydrogenation) to improve its properties, due to the content of unsaturated hydrocarbons ^{[43][44][45]}.

It should also be considered that even well-sorted waste for fuel production (PE, PP and PS) may contain contaminants in the form of other plastics, e.g., polyvinyl chloride (PVC). The presence of this plastic will result in the formation of chlorine compounds (including hydrogen chloride) during the recycling process. These compounds will have a corrosive effect on apparatus components and poison the catalyst ^[46]. Since the proportion of PVC admixture is generally low, in order to eliminate the harmful effect of chlorine, chlorine-binding compounds such as NaHCO₃, CaO, CaCO₃, Na₂CO₃, Ca(OH)₂ can be fed to the reactor with the feedstock ^[47].

Plastics, among the additives, may contain flame-retardant compounds. These additives are usually bromine or chlorine derivatives, e.g., chlorinated paraffins, boric acid; polybrominated diphenyl ethers, tetrabromobisphenol, tris(2-chloroethyl)phosphate, hexabromocyclohexane ^[48]. These compounds are generally not added to packaging, which has the largest share in the plastic waste stream, while they may be included in EPS waste from construction. The degradation of bromine compounds can lead to the formation of organo-bromine compounds and bromine hydrogen, whose action is similar to that of chlorine compounds. HBr can be disposed of in a manner analogous to hydrogen chloride. Another method to inactivate bromine compounds is the use of suitable catalysts

during pyrolysis, e.g., metal oxides (e.g., ZnO, Fe₂O₃, La₂O₃, Cu₂O, Sb₂O₃, Al₂O₃, [49][50][51]), calcium and ironbased composites [52][53], zeolites, molecular sieves, red mud, or natural zeolites [54][55].

The reactor also plays an important role in the pyrolysis process ^[56]. The type of reactor used affects the process control, mass and heat exchange, mixing, liquid and gas phase flows, retention time, etc. There are many ways to classify pyrolysis reactors. According to the method of carrying out the process, there are batch reactors, semi-batch reactors and reactors with continuous flow. In turn, due to their construction, fixed bed, fluidized bed (bubbling and circulating) and conical spouted bed reactors (CSBR) can be distinguished, among others. Another criterion for the division may be, for example, the method of heat energy supply: to the interior via an inert gas or material, through the external walls via flue gas or electricity, or via internal heating elements ^{[57][58][59]}. Examples of studies focused on plastic waste pyrolysis using different reactors are shown in **Table 1**. The main advantages and disadvantages of pyrolysis reactors are presented in **Table 2**.

Type of Reactor	Type of Feedstock Materials	References
	PS, PE, PP and PET	[<u>54]</u>
Batch reactor	HDPE	[25]
	Plastic medical wastes	[<u>60</u>]
	PS, HDPE, LDPE, PP	[<u>61]</u>
Semi-batch reactor	PE	[<u>62]</u>
	Polyolefin	[63]
Continuous reactor	Plastic waste from landfill	[64]
	HDPE	[<u>65</u>]
Fixed bed reactor	PP, HDPE, LDPE	[<u>66][67]</u>

Table 1. Examples of reactors for polymer pyrolysis.

Type of R	eactor	Type of Feedsto	ck Materials	References
		HDPI	E	[<u>68]</u>
		LDPE	E	[<u>69]</u>
Fluidized bed react	or	PP, P	E	[70]
		PE		[71]
		HDPE, LDF	PE, PP	[72]
CSBR		HDPE, LDPE, PP, F	PS, PET, PMMA	[<u>73</u>]
		PE	I	[<u>27][39][57</u>]
Type of Reactor		Advantages	Disadvantag	jes
Batch reactor	High converse	sions of feedstock.	 Variability of product fr batch; High labor costs per base Problem with adoption production. 	om batch to atch; for large-scale
Semi-batch reactor	 Possibility of reagents and Flexibility of 	simultaneous addition of d removal of the product; adding reactants over time.	 High labor costs per base Problems with adoption production. 	atch; n for large-scale
Fixed-bed reactor	Simple and i	reliable construction.	 Keeping the size and s material particles relati Limited accessibility of surface area during the 	shape of the raw ively constant; the catalyst e reaction;

Type of Reactor	Advantages	Disadvantages
		 Low heating rate; Non-uniform temperature inside the polymeric material.
Fluidized-bed reactor	 Better access to the catalyst than in the fixed-bed reactor; Large reaction surface; Higher degree of cracking than in the fixed-bed reactor; Uniform temperature distribution; Excellent mixing properties; Better consistency of the process conditions; Less frequent loading of raw material and the associated process shutdown; Easy catalyst regeneration without unloading. 	 Small particle size of the raw material; Problem with separating char from bed material; It is easy to defluidize bed when melted plastic sticks on it.
Rotary kiln reactor	 More efficient in heating raw materials than a fixed-bed reactor; Good mixing of raw material; Yields uniform pyrolytic products; Possibility of adjusting the residence time; Simple maintenance; 	 Slow heating rate provides a significant amount of char to the products; Mainly applicable to conventional pyrolysis.

Type of Reactor	Advantages	Disadvantages
	• Possibility of feeding the raw material with different particle sizes.	
Stirred tank reactor	 Good temperature control; Easy to construct and operate; Uniform heat distribution. 	 Tendency to build up char deposits on the walls; Low conversion rate per volume; Poor agitation; Require frequent maintenance.
Conical spouted bed reactor	 Good mixing; Possibility of using raw material with a large particle size distribution; High heat transfer; Protects against defluidization caused by melted plastic; Relatively simple construction; Extremely short residence time of the polymers in reactor avoids secondary reactions; Lower pressure drop in comparison with other fluidized bed reactor. 	 Problem with catalyst feeding; Problem with catalyst entrainment; Deposition of products on the catalyst; The reactor requires many pumps (high operating costs).



Figure 3. Pyrolytic methods of plastic conversion.

References

- World Energy Outlook 2021. Report of International Energy Agency. October 2021. Available online: https://iea.blob.core.windows.net/assets/4ed140c1-c3f3-4fd9-acae-789a4e14a23c/WorldEnergyOutlook2021.pdf (accessed on 10 January 2022).
- Short-Term Energy Outlook. Report of U.S. Energy Information Adsministration. December 2021. Available online: https://www.eia.gov/outlooks/steo/pdf/steo_full.pdf (accessed on 10 January 2022).
- 3. Umar, M.; Farid, S.; Naeem, M.A. Time-frequency connectedness among clean-energy stocks and fossil fuel markets: Comparison between financial, oil and pandemic crisis. Energy 2022, 240, 122702.
- Tang, C.; Aruga, K. Relationships among the Fossil Fuel and Financial Markets during the COVID-19 Pandemic: Evidence from Bayesian DCC-MGARCH Models. Sustainability 2022, 14, 51.
- 5. Strielkowski, W.; Civín, L.; Tarkhanova, E.; Tvaronavičienė, M.; Petrenko, Y. Renewable Energy in the Sustainable Development of Electrical Power Sector: A Review. Energies 2021, 14, 8240.
- Pisciotta, M.; Pilorgé, H.; Feldmann, J.; Jacobson, R.; Davids, J.; Swett, S.; Sasso, Z.; Wilcox, J. Current state of industrial heating and opportunities for decarbonization. Prog. Energy Combust. Sci. 2022, 91, 100982.

- 7. Special Report. Global Warming of 1.5 °C. Available online: https://www.ipcc.ch/sr15/ (accessed on 10 January 2022).
- Shooshtarian, S.; Maqsood, T.; Caldera, S.; Ryley, T. Transformation towards a circular economy in the Australian construction and demolition waste management system. Sustain. Prod. Consum. 2022, 30, 89–106.
- 9. Sikdar, S. Circular economy: Is there anything new in this concept? Clean Technol. Environ. Policy 2019, 21, 1173–1175.
- Plastics—The Facts 2021—Report Plastics Europe. The Association of Plastics Manufacturers in Europe. Available online: https://plasticseurope.org/pl/knowledge-hub/plastics-the-facts-2021/ (accessed on 13 January 2022).
- Menyuka, N.N.; Sibanda, M.; Bob, U. Perceptions of the Challenges and Opportunities of Utilising Organic Waste through Urban Agriculture in the Durban South Basin. Int. J. Environ. Res. Public Health 2020, 17, 1158.
- 12. Jha, K.K.; Kannan, T.T.M. Recycling of plastic waste into fuel by pyrolysis—A review. Mater. Today Proc. 2021, 37, 3718–3720.
- 13. Kang, M.J.; Yu, H.J.; Jegal, J.; Kim, H.S.; Cha, H.G. Depolymerization of PET into terephthalic acid in neutral media catalyzed by the ZSM-5 acidic catalyst. Chem. Eng. J. 2020, 398, 125655.
- 14. Balema, V.P.; Hlova, I.Z.; Carnahan, S.L.; Seyedi, M.; Dolotko, O.; Rossini, A.J.; Luzinov, I. Depolymerization of polystyrene under ambient conditions. New J. Chem. 2021, 45, 2935–2938.
- 15. Dos Santos, P.B.; da Silva Ribeiro, H.J.; Ferreira, A.C.; Ferreira, C.C.; Bernar, L.P.; da Costa Assunção, F.P.; de Castro, D.A.R.; Santos, M.C.; Duvoisin, S.; Borges, L.E.P.; et al. Process Analysis of PMMA-Based Dental Resins Residues Depolymerization: Optimization of Reaction Time and Temperature. Energies 2022, 15, 91.
- 16. Lahtela, V.; Hyvärinen, M.; Kärki, T. Composition of Plastic Fractions in Waste Streams: Toward More Efficient Recycling and Utilization. Polymers 2019, 11, 69.
- 17. Jubinville, D.; Esmizadeh, E.; Saikrishnan, S.; Tzoganakis, C.; Mekonnen, T. A comprehensive review of global production and recycling methods of polyolefin (PO) based products and their post-recycling applications. Sustain. Mater. Technol. 2020, 25, e00188.
- Shen, L.; Worrell, E. Chapter 13—Plastic Recycling. In Handbook of Recycling. State-of-the-Art for Practitioners, Analysts, and Scientists; Worrell, E., Reuter, M.A., Eds.; Elsevier: Amsterdam, The Netherlands, 2014; pp. 179–190.
- 19. Singh, N.; Hui, D.; Singh, R.; Ahuja, I.P.S.; Feo, L.; Fraternali, F. Recycling of plastic solid waste: A state of art review and future applications. Compos. Part B Eng. 2017, 115, 409–422.

- 20. Ignatyev, I.A.; Thielemans, W.; Vander Beke, B. Recycling of Polymers: A Review. ChemSusChem 2014, 7, 1579–1593.
- 21. Schwabl, D.; Bauer, M.; Lehner, M. Advancing Plastic Recycling by Wet-Mechanical Processing of Mixed Waste Fractions. Processes 2021, 9, 493.
- Ragaert, K.; Huysveld, S.; Vyncke, G.; Hubo, S.; Veelaert, L.; Dewulf, J.; Du Bois, E. Design from recycling: A complex mixed plastic waste case study. Resour. Conserv. Recycl. 2020, 155, 104646.
- 23. Solis, M.; Silveira, S. Technologies for chemical recycling of household plastics—A technical review and TRL assessment. Waste Manag. 2020, 105, 128–138.
- 24. Li, N.; Liu, H.X.; Cheng, Z.N.; Yan, B.B.; Chen, G.Y.; Wang, S.B. Conversion of plastic waste into fuels: A critical review. J. Hazard. Mater. 2022, 424, 127460.
- 25. Palos, R.; Gutiérrez, A.; Vela, F.J.; Maña, J.A.; Hita, I.; Asueta, A.; Arnaiz, S.; Arandes, J.M.; Bilbao, J. Assessing the potential of the recycled plastic slow pyrolysis for the production of streams attractive for refineries. J. Anal. Appl. Pyrolysis 2019, 142, 104668.
- 26. Prajapati, R.; Kohli, K.; Maity, S.K.; Sharma, B.K. Potential Chemicals from Plastic Wastes. Molecules 2021, 26, 3175.
- 27. Gebre, S.H.; Sendeku, M.G.; Bahri, M. Recent Trends in the Pyrolysis of Non-Degradable Waste Plastics. ChemistryOpen 2021, 10, 1202–1226.
- 28. Ma, H.; Cao, Y.; Lu, X.; Ding, Z.; Zhou, W. Review of Typical Municipal Solid Waste Disposal Status and Energy Technology. Energy Procedia 2016, 88, 589–594.
- 29. Arandes, J.M.; Torre, I.; Castaño, P.; Olazar, M.; Bilbao, J. Catalytic Cracking of Waxes Produced by the Fast Pyrolysis of Polyolefins. Energy Fuels 2007, 21, 561–569.
- Rodríguez, E.; Gutiérrez, A.; Palos, R.; Vela, F.J.; Arandes, J.M.; Bilbao, J. Fuel production by cracking of polyolefins pyrolysis waxes under fluid catalytic cracking (FCC) operating conditions. Waste Manag. 2019, 93, 162–172.
- 31. Fivga, A.; Dimitriou, I. Pyrolysis of plastic waste for production of heavy fuel substitute: A technoeconomic assessment. Energy 2018, 149, 865–874.
- Fekhar, B.; Zsinka, V.; Miskolczi, N. Value added hydrocarbons obtained by pyrolysis of contaminated waste plastics in horizontal tubular reactor: In situ upgrading of the products by chlorine capture. J. Clean. Prod. 2019, 241, 118166.
- 33. Kaminsky, W. Chemical recycling of plastics by fluidized bed pyrolysis. Fuel Commun. 2021, 8, 100023.

- Syamsiro, M.; Saptoadi, H.; Norsujianto, T.; Noviasri, P.; Cheng, S.; Alimuddin, Z.; Yoshikawa, K. Fuel Oil Production from Municipal Plastic Wastes in Sequential Pyrolysis and Catalytic Reforming Reactors. Energy Procedia 2014, 47, 180–188.
- 35. Bezergianni, S.; Dimitriadis, A.; Faussone, G.C.; Karonis, D. Alternative Diesel from Waste Plastics. Energies 2017, 10, 1750.
- 36. Nanda, S.; Berruti, F. Thermochemical conversion of plastic waste to fuels: A review. Environ. Chem. Lett. 2021, 19, 123–148.
- Marcilla, A.; Beltrán, M.I.; Navarro, R. Thermal and catalytic pyrolysis of polyethylene over HZSM5 and HUSY zeolites in a batch reactor under dynamic conditions. Appl. Catal. B Environ. 2009, 86, 78–86.
- 38. Fahim, I.; Mohsen, O.; ElKayaly, D. Production of Fuel from Plastic Waste: A Feasible Business. Polymers 2021, 13, 915.
- 39. Sharuddin, S.D.A.; Abnisa, F.; Daud, W.M.A.W.; Aroua, M.K. A review on pyrolysis of plastic wastes. Energy Convers. Manag. 2016, 115, 308–326.
- 40. Papari, S.; Bamdad, H.; Berruti, F. Pyrolytic Conversion of Plastic Waste to Value-Added Products and Fuels: A Review. Materials 2021, 14, 2586.
- 41. Papari, S.; Hawboldt, K. Development and Validation of a Process Model to DescribePyrolysis of Forestry Residues in an Auger Reactor. Energy Fuels 2017, 31, 10833–10841.
- 42. Suresh, A.; Alagusundaram, A.; Kumar, P.S.; Vo, D.V.N.; Christopher, F.C.; Balaji, B.; Viswanathan, V.; Sankar, S. Microwave pyrolysis of coal, biomass and plastic waste: A review. Environ. Chem. Lett. 2021, 19, 3609–3629.
- Miandad, R.; Nizami, A.S.; Rehan, M.; Barakat, M.A.; Khan, M.I.; Mustafa, A.; Ismail, I.M.I.; Murphy, J.D. Influence of temperature and reaction time on the conversion of polystyrene waste to pyrolysis liquid oil. Waste Manag. 2016, 58, 250–259.
- 44. Diaz-Silvarrey, L.S.; Zhang, K.; Phan, A.N. Monomer recovery through advanced pyrolysis of waste high density polyethylene (HDPE). Green Chem. 2018, 20, 1813–1823.
- 45. Miandad, R.; Barakat, M.A.; Aburiazaiza, A.S.; Rehan, M.; Ismail, I.M.I.; Nizami, A.S. Effect of plastic waste types on pyrolysis liquid oil. Int. Biodeterior. Biodegrad. 2017, 119, 239–252.
- 46. Grabda, M.; Oleszek, S.; Shibata, E.; Nakamura, T. Study on simultaneous recycling of EAF dust and plastic waste containing TBBPA. J. Hazard. Mater. 2014, 278, 97–106.
- Fekhar, B.; Gombor, L.; Miskolczi, N. Pyrolysis of chlorine contaminated municipal plastic waste: In-situ upgrading of pyrolysis oils by Ni/ZSM-5, Ni/SAPO-11, red mud and Ca(OH)2 containing catalysts. J. Energy Inst. 2019, 92, 1270–1283.

- 48. Hahladakis, J.N.; Velis, C.A.; Weber, R.; Iacovidou, E.; Purnell, P. An overview of chemical additives present in plastics: Migration, release, fate and environmental impact during their use, disposal and recycling. J. Hazard. Mater. 2018, 344, 179–199.
- 49. Oleszek, S.; Kumagai, S.; Grabda, M.; Shiota, K.; Yoshioka, T.; Takaoka, M. Mitigation of brominecontaining products during pyrolysis of polycarbonate-based tetrabromobisphenol A in the presence of copper(I) oxide. J. Hazard. Mater. 2021, 409, 124972.
- 50. Ma, C.; Yu, J.; Wang, B.; Song, Z.; Xiang, J.; Hu, S.; Su, S.; Sun, L. Chemical recycling of brominated flame retarded plastics from e-waste for clean fuels production: A review. Renew. Sustain. Energy Rev. 2016, 61, 433–450.
- 51. Terakado, O.; Ohhashi, R.; Hirasawa, M. Bromine fixation by metal oxide in pyrolysis of printed circuit board containing brominated flame retardant. J. Anal. Appl. Pyrolysis 2013, 103, 216–221.
- 52. Jung, S.H.; Kim, S.J.; Kim, J.S. Thermal degradation of acrylonitrile–butadiene–styrene (ABS) containing flame retardants using a fluidized bed reactor: The effects of Ca-based additives on halogen removal. Fuel Process Technol. 2012, 96, 265–270.
- 53. Brebu, M.; Bhaskar, T.; Murai, K.; Muto, A.; Sakata, Y.; Uddin, M.A. Removal of nitrogen, bromine, and chlorine from PP/PE/PS/PVC/ABS–Br pyrolysis liquid products using Fe- and Ca-based catalysts. Polym. Degrad. Stab. 2005, 87, 225–230.
- 54. Manžuch, Z.; Akelytė, R.; Camboni, M.; Carlander, D. Chemical Recycling of Polymeric Materials from Waste in the Circular Economy. RPA Europe Final Report. August 2021. Available online: https://echa.europa.eu/documents/10162/1459379/chem_recycling_final_report_en.pdf/887c4182-8327-e197-0bc4-17a5d608de6e?t=1636708465520, (accessed on 14 January 2022).
- 55. Charitopoulou, M.A.; Kalogiannis, K.G.; Lappas, A.A.; Achilias, D.S. Novel trends in the thermochemical recycling of plastics from WEEE containing brominated flame retardants. Environ. Sci. Pollut. Res. 2021, 28, 59190–59213.
- 56. Al-Salem, S.M.; Antelava, A.; Constantinou, A.; Manos, G.; Dutta, A. A review on thermal and catalytic pyrolysis of plastic solid waste (PSW). J. Environ. Manag. 2017, 197, 177–198.
- 57. Wong, S.L.; Ngadi, N.; Abdullah, T.A.T.; Inuwa, I.M. Current state and future prospects of plastic waste as source of fuel: A review. Renew. Sustain. Energy Rev. 2015, 50, 1167–1180.
- 58. Lewandowski, W.M.; Januszewicz, K.; Kosakowski, W. Efficiency and proportions of waste tyre pyrolysis products depending on the reactor type—A review. J. Anal. Appl. Pyrolysis 2019, 140, 25–53.
- Uddin, M.N.; Techato, K.; Taweekun, J.; Rahman, M.M.; Rasul, M.G.; Mahlia, T.M.I.; Ashrafur, S.M. An Overview of Recent Developments in Biomass Pyrolysis Technologies. Energies 2018, 11, 3115.

- 60. Rasul, S.B.; Som, U.; Hossain, S.; Rahman, W. Liquid fuel oil produced from plastic based medical wastes by thermal cracking. Sci. Rep. 2021, 11, 17048.
- 61. Lee, K.H.; Shin, D.H. Characteristics of liquid product from the pyrolysis of waste plastic mixture at low and high temperatures: Influence of lapse time of reaction. Waste Manag. 2007, 27, 168–176.
- 62. Pan, R.; Martins, M.F.; Debenest, G. Pyrolysis of waste polyethylene in a semi-batch reactor to produce liquid fuel: Optimization of operating conditions. Energy Convers. Manag. 2021, 237, 114114.
- 63. Torres, D.; Jiang, Y.; Sanchez Monsalve, D.A.; Leeke, G.A. Chlorine removal from the pyrolysis of urban polyolefinic waste in a semi-batch reactor. J. Environ. Chem. Eng. 2021, 9, 104920.
- 64. Al-Salem, S.M.; Yang, Y.; Wang, J.; Leeke, G.A. Pyro-Oil and Wax Recovery from Reclaimed Plastic Waste in a Continuous Auger Pyrolysis Reactor. Energies 2020, 13, 2040.
- Zhou, N.; Dai, L.; Lv, Y.; Li, H.; Deng, W.; Guo, F.; Chen, P.; Lei, H.; Ruan, R. Catalytic pyrolysis of plastic wastes in a continuous microwave assisted pyrolysis system for fuel production. Chem. Eng. J. 2021, 418, 129412.
- 66. Tulashie, S.K.; Boadu, E.K.; Dapaah, S. Plastic waste to fuel via pyrolysis: A key way to solving the severe plastic waste problem in Ghana. Therm. Sci. Eng. Prog. 2019, 11, 417–424.
- 67. Papuga, S.V.; Gvero, P.M.; Vukić, L.M. Temperature and time influence on the waste plastics pyrolysis in the fixed bed reactor. Therm. Sci. 2016, 20, 731–741.
- Li, H.; Mašek, O.; Harper, A.; Ocone, R. Kinetic study of pyrolysis of high-density polyethylene (HDPE) waste at different bed thickness in a fixed bed reactor. Can. J. Chem. Eng. 2021, 99, 1733–1744.
- 69. Milne, B.J.; Behie, L.A.; Berruti, F. Recycling of waste plastics by ultrapyrolysis using an internally circulating fluidized bed reactor. J. Anal. Appl. Pyrolysis 1999, 51, 157–166.
- Jung, S.H.; Cho, M.H.; Kang, B.S.; Kim, J.S. Pyrolysis of a fraction of waste polypropylene and polyethylene for the recovery of BTX aromatics using a fluidized bed reactor. Fuel Process. Technol. 2010, 91, 277–284.
- 71. Mastral, J.F.; Berrueco, C.; Ceamanos, J. Modelling of the pyrolysis of high density polyethylene: Product distribution in a fluidized bed reactor. J. Anal. Appl. Pyrolysis 2007, 79, 313–322.
- 72. Arabiourrutia, M.; Elordi, G.; Lopez, G.; Borsella, E.; Bilbao, J.; Olazar, M. Characterization of the waxes obtained by the pyrolysis of polyolefin plastics in a conical spouted bed reactor. J. Anal. Appl. Pyrolysis 2012, 94, 230–237.
- 73. Orozco, S.; Alvarez, J.; Lopez, G.; Artetxe, M.; Bilbao, J.; Olazar, M. Pyrolysis of plastic wastes in a fountain confined conical spouted bed reactor: Determination of stable operating conditions.

Energy Convers. Manag. 2021, 229, 113768.

74. Artetxe, M.; Lopez, G.; Elordi, G.; Amutio, M.; Bilbao, J.; Olazar, M. Production of Light Olefins from Polyethylene in a Two-Step Process: Pyrolysis in a Conical Spouted Bed and Downstream High-Temperature Thermal Cracking. Ind. Eng. Chem. Res. 2012, 51, 13915–13923.

Retrieved from https://encyclopedia.pub/entry/history/show/84141