### **Paper-Based Analytical Devices for Colorimetric**

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The so-called paper-based analytical devices (PADs) have arisen as an efficient, affordable, user-friendly, rapid, and equipment-free technology that is available to citizens. The development of PADs in areas such as clinical diagnostics, food safety and environmental monitoring, etc., as well as fabrication methods, target analytes and analytical performance, has been extensively reviewed during the last decade, with the scientific community showing great interest toward these appealing analytical approaches.

Keywords: paper-based analytical devices ; mercury ; nanomaterials

### 1. Introduction

Hg can reach aquatic ecosystems through point-source discharges or atmospheric deposition. Thus, volcanic eruptions and the solubilization of rocks, soils and sediments are among the most relevant natural sources <sup>[1]</sup>. Anthropogenic sources such as small-scale gold mining, the combustion of solid fuels (coal, lignite, wood), chlor-alkali, paper, paint and pharmaceutical industries, dental implants, agriculture products (germicides, pesticides, etc.); although mostly restricted in many countries, they still contribute to increasing the Hg levels in the environment <sup>[2]</sup>. Therefore, stringent analytical controls are needed to assess the contamination of environmental samples with Hg.

Hg can be found in different environmental compartments as a variety of species; each one has different behavior, and hence, toxicological properties, bioavailability and environmental impact depend on its physicochemical forms (i.e., speciation). Thus, in natural waters, the main forms in which Hg can be present are elemental mercury (Hg 0), inorganic mercury (Hg 2+) and organic mercury, i.e., CH 3-Hg + and (CH 3) 2Hg. Biomagnification of Hg through the food chain may occur as a result of the high hydrophobicity of organic Hg species. In this way, Hg can accumulate in some fish by a factor of ca. 10 6 in respect to the concentration levels in the aquatic environment <sup>[3]</sup>.

For the determination of Hg at the (ultra)trace level, conventional instrumentation is typically used in central labs on a routine basis, such as cold vapor-atomic absorption spectrometry (CV-AAS) <sup>[4]</sup>, cold vapor-atomic fluorescence spectrometry (CV-AFS) <sup>[5]</sup>, electrothermal atomic absorption spectrometry (ETAAS) <sup>[6]</sup>, inductively coupled plasma-mass spectrometry (ICP-MS) <sup>[7]</sup> and total reflection X-ray fluorescence (TXRF) <sup>[8]</sup>. While these techniques provide adequate sensitivity and precision, they require suitable sampling, preservation procedures, sample pretreatment and a fully controlled laboratory environment, which makes it difficult to extend their application for on-field analysis <sup>[9]</sup>. In addition, problems may arise in the sampling and sample preparation procedures prior to the determination of Hg at the (ultra)trace level by conventional analytical techniques, which can lead to systematic errors and unacceptable analytical uncertainties <sup>[3]</sup>.

In recent years, several trends have emerged concerning the analytical control of environmental pollutants, such as a remarkable increase in the miniaturization, portability and greenness of analytical approaches, thus facilitating on-site measurements <sup>[10]</sup>. The latter possibility is particularly interesting, since it could allow real time measurements without the need for preservation, transport and sample storing prior to analysis by a conventional technique. Further appealing features include the possibility of performing temporally and spatially discriminated analysis and the access to remote sites so that the source of pollutants, their distribution and environmental impact can be more easily assessed <sup>[11]</sup>.

#### 2. Development of Paper-Based Analytical Devices for the Detection of Mercury

Lab-on-chip (LOC) technologies have emerged as miniaturized, low cost and fast analytical approaches allowing a decrease in sample, reagents and energy consumption through the integration of typical stages of bench-scale laboratories within a single device <sup>[12]</sup>. From the standpoint of green chemistry, the use of cellulose instead of typical substrates employed in LOC systems such as polymers, silicon or glass represent a significant step forward. Cellulose-

based materials have been established in the last years as efficient, versatile and universal biopolymers for the design of novel microscale analytical systems <sup>[13]</sup> (**Figure 1**). As compared to other scaffolds used for building sensors and microfluidic devices, cellulose is a biodegradable, biocompatible, hydrophilic and highly porous material. In addition, it possesses high capillarity, and a large variety of recognition elements can be immobilized for sensing. When used along the widespread colorimetric transduction, white color is excellent to achieve good analytical performance <sup>[14]</sup>.



Figure 1. Conceptual scheme showing the application of cellulose for the detection of Hg.

The so-called paper-based analytical devices (PADs) have arisen as an efficient, affordable, user-friendly, rapid, and equipment-free technology that is available to citizens. The development of PADs in areas such as clinical diagnostics, food safety and environmental monitoring, etc., as well as fabrication methods, target analytes and analytical performance, has been extensively reviewed during the last decade [15][16][17][18][19][20][21][22][23][24][25][26][27][28][29], with the scientific community showing great interest toward these appealing analytical approaches.

Under the general term 'paper-based analytical devices' (PADs), two systems can be distinguished, i.e., microfluidic paper-based analytical devices' ( $\mu$ -PAD), where a fluidic network is built in the paper substrate, and 'paper-based assay devices', also known as 'paper-based sensors' or 'spot tests', where the sample is directly deposited onto the paper surface. First systems, introduced by Whitesides for the first time <sup>[30]</sup>, include different configurations, such as two-dimensional (2D), three-dimensional (3D) and distance-based devices. In these microfluidic devices, the sample and reagents are transported to the detection zone by capillarity. Second designs derive from the classical qualitative analysis tests, where the detection of inorganic cations and anions could be performed on filter paper using suitable colorimetric and fluorescent reagents <sup>[31]</sup>. In paper-based assay devices, the sample comes directly into contact with the receptor, which remains stationary on the cellulose scaffold.

In this review, we provide an overview on the state of the art of PADs for the detection of Hg in environmental samples, their main shortcomings and future prospects.

## 3. Paper-Based Sensors Integrated with Organic Chromogenic/Fluorogenic Receptors for Hg Detection

Several chromogenic/fluorogenic reagents have been used as recognition elements for the detection of Hg(II) in both paper-based sensors [32][33][34][35][36][37][38] and  $\mu$ -PADs [39][40][41][42] (**Table 1**). In a few cases, multiplexed systems for the detection of other metal ions have been reported [41][42]. An array of paper strips has also been designed for the detection of several metals, including Hg [43]. Environmental samples analyzed mostly include several types of waters, yet applications to biological samples, soils and creams have also been described [32][39][40]. With some exceptions where inorganic chromogenic species are involved [32][40], the most reported applications use organic chromogenic reagents for analyte recognition.

**Table 1.** Applications of chromogenic/fluorogenic reagents as receptors in PADs (µ-PADs and paper-based sensors) for the detection of Hg(II).

Material	Type of PAD	Recognition Element	Signal Readout	Sample/Matrix	LOD (ppb)	Ref.
3 MM Whatman chromatography paper	Paper-based sensor	Cul	Scanner	Fish	7 (ng/g)	[ <u>32</u> ]
Cellulose	Paper-based sensor	bis(ferrocenyl) azine	Naked eye	Aqueous media	104	[ <u>33]</u>

Material	Type of PAD	Recognition Element	Signal Readout	Sample/Matrix	LOD (ppb)	Ref.
Porous silica matrix onto cellulose	Paper-based sensor *	Rhodamine B thiolactone	Flatbed scanner and naked eye	Water	0.24 (Scanner)	[ <u>34</u> ]
Filter paper	Paper-based sensor	Rhodamine appended vinyl ether	Naked eye	Drinking water Tap water	27.2 (in solution) 10 <sup>4</sup> (paper strip)	<u>[35]</u>
Whatman paper	Paper-based sensor	Ir complex (Phosphorescent)	Naked eye		3.56 (fluorimetry)	[ <u>36]</u>
Cellulose paper	Paper-based sensor (Hg, I, Zn)	Calix[4]arene (fluorescent)	Digital camera (UV irradiation)	Wastewater	0.58 (fluorimetry)	[ <u>37]</u>
Filter paper	Paper-based sensor	Tetrahydrophenazine- based Fluorophore	Digital camera		8 × 10 <sup>3</sup> (neutral pH) 3 × 10 <sup>3</sup> (pH 1.6– 2.3)	<u>[38]</u>
Filter paper	μ-PAD	Dithizone	Naked eye	Whitening cream	930	[ <u>39]</u>
Whatman No. 4 filter paper	μ-PAD	Hgl4 <sup>2-</sup> complex	Digital camera	Contaminated soil and water	2 × 10 <sup>4</sup>	<u>[40]</u>
Whatman grades No. 1 and 4	µ-PAD (Hg, Pb, Cr, Cu, Fe)	Three indicators (ligands)	Digital camera	Waters	20	<u>[41]</u>
Whatman No. 1 paper	µ-PAD (Cu, Co, Ni, Mn, Hg)	Dithizone (for Hg)	Scanner	Drinking, pond and tap water	200 (scanner) <i>ca.</i> 10 <sup>4</sup> (naked eye)	[42]
Whatman grade No. 1 filter paper	Array paper strip (for Hg, Ag, Cu)	5 indicators (18 formulations)	Flatbed scanner	Pond water	38 (Hg)	[43]

\* PADs providing an LOD equal or below the maximum contaminant level fixed by US EPA.

While naked eye detection is carried out in many PADs, devices related to information and communication technologies (ICTs) such as digital cameras, scanners, smartphones, etc., have been mostly used for capturing images on PADs. Further image processing is employed for measuring color intensity. LODs at the ppm level are generally reported for many applications of PADs concerning Hg detection, with the exception of approaches involving any kind of preconcentration (e.g., <sup>[34]</sup>). In a significant number of papers, LODs corresponding to the use of receptors in a solution followed by detection with a conventional instrument are provided (e.g., <sup>[36][37]</sup>).

Patil and Das <sup>[35]</sup> described a selective colorimetric and fluorometric chemosensor based on a rhodamine appended vinyl ether (RDV) probe for Hg(II) recognition. Paper strips were employed by immersing filter paper into a RDV solution. Although an LOD of 136 nM was obtained using a solution assay, the paper strip was useful for Hg detection at the ppm level (above ca. 10 ppm).

The immobilization of an infrared fluorescence protein (IFP) and its chromophore biliverdin (BV) has been applied by Gu et al. <sup>[44]</sup> for Hg(II) detection. An LOD of less than 50 nM was achieved. The IFP/BV sensor can serve as a tool for the detection of Hg in living organisms or tissues. A protein-hydrogel-based paper assay was also used for the immobilization of IFP onto paper strips for detection of Hg(II). Enrichment by multiple addition/drying steps onto the paper strip allows detection at the 20 nM level.

# 4. Paper-Based Analytical Devices Integrated with Nanomaterials as Receptors for Hg Detection

When the light of appropriate frequency interacts with some metal nanoparticles (e.g., Au, Ag, Cu), a collective oscillation of electrons at their conduction bands occurs, which is the basis for the surface plasmon resonance (SPR) phenomenon <sup>[45]</sup>. When the dimensions of metal nanoparticles are lesser than the radiation wavelength, the phenomenon is known as

'localized surface plasmon resonance' (LSPR). Absorption of radiation takes place when light has the same frequency as oscillations. The localized surface plasmon resonance (LSPR) absorption bands are characteristics of the metal involved in the colloidal solution, i.e., it depends on size, shape, interparticle distance, composition of the nanoparticles and refractive index of the surrounding medium. Thus, the colors displayed by colloidal solutions of AuNPs, AgNPs and CuNPs are pink, yellow and red, respectively. More interestingly, these NPs possess much higher molar extinction coefficients as compared to chromogenic agents. The molar extinction coefficients corresponding to the LSPR absorption bands of AuNPs and AgNPs are 10 8 and 10 10 M -1 cm -1, respectively. Typically, the wavelength of the LSPR band is largely affected by the size and chemical environment surrounding the nanoparticles, such as the presence of capping agents, formation of amalgams, species adsorbed, etc. Noble metal nanoparticles have been widely applied for the detection of metal ions [46].

A paper-based sensor was developed by Apilux et al. <sup>[47]</sup> for the detection of Hg(II) in waters using AgNPs and silver nanoplates (AgNPls). The color change of AgNPls on a paper test in the presence of Hg(II) can be monitored by the naked eye. A quantitative assay can be accomplished following image capture by a digital camera along with an image processing software to yield an LOD of 0.12 ppm. Upon the accumulation of Hg on paper through multiple applications of 2  $\mu$ L, an LOD of 2 ppb can be achieved. The color change of AgNPs and AgNPls can be ascribed to changes in size and shape. A sensing mechanism based on the redox reaction between Hg(II) and AgNPls was proposed.

An inkjet-printed paper-based colorimetric sensor with AgNPs, along with a smartphone and RGB color detection, was developed by Shrivas et al.  $^{[48]}$  (**Figure 2** B). A color change from yellow to colorless was observed in the presence of Hg(II). A reaction mechanism responsible for the color change was proposed as a result of the interaction of Hg(II) and a PVP stabilizing agent employed as a capping agent for AgNPs, and an oxidation of Ag 0 to Ag + . An LOD of 10 ppb was obtained.



**Figure 2.** (A) Scheme of the PAD preparation and digital image acquisition for Hg detection using AgNPs as colorimetric probe <sup>[49]</sup>. (B) Scheme for inkjet-printing of PVP-AgNPs on paper; (1) synthesis of PVP-AgNPs; (2) Inkjet-printing on Whatman filter papers; (3) deposition of sample solution containing Hg(II) on the PAD; (4) image capture using a smartphone; (5) measurement of color intensity; (6) standard calibration curve <sup>[48]</sup>. **Figure 2**B is reproduced with permission of Elsevier <sup>[48]</sup>.

Metal nanoclusters (NCs) made of Au, Ag, Cu, etc., with a size of less than 2 nm, does do not undergo the SPR effect, unlike metal nanoparticles (NPs), but they possess strong luminescence <sup>[50]</sup>.

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