Respirable Coal Mine Dust Exposure

Subjects: Mineralogy

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Respirable coal mine dust (RCMD) exposure is associated with black lung and silicosis diseases in underground miners. Although only RCMD mass and silica concentrations are regulated, it is possible that particle size, surface area, and other chemical constituents also contribute to its adverse health effects. This review summarizes measurement technologies for RCMD mass concentrations, morphology, size distributions, and chemical compositions, with examples from published efforts where these methods have been applied. Some state-of-the-art technologies presented in this paper have not been certified as intrinsically safe, and caution should be exerted for their use in explosive environments. RCMD mass concentrations are most often obtained by filter sampling followed by gravimetric analysis, but recent requirements for real-time monitoring by continuous personal dust monitors (CPDM) enable quicker exposure risk assessments. Emerging low-cost photometers provide an opportunity for a wider deployment of real-time exposure assessment. Particle size distributions can be determined by microscopy, cascade impactors, aerodynamic spectrometers, optical particle counters, and electrical mobility analyzers, each with unique advantages and limitations. Different filter media are required to collect integrated samples over working shifts for comprehensive chemical analysis. Teflon membrane filters are used for mass by gravimetry, elements by energy dispersive X-ray fluorescence, rare-earth elements by inductively coupled plasmamass spectrometry and mineralogy by X-ray diffraction. Quartz fiber filters are analyzed for organic, elemental, and brown carbon by thermal/optical methods and non-polar organics by thermal desorption-gas chromatography-mass spectrometry. Polycarbonate-membrane filters are analyzed for morphology and elements by scanning electron microscopy (SEM) with energy dispersive X-ray, and quartz content by Fourier-transform infrared spectroscopy and Raman spectroscopy.

Keywords: respirable coal mine dust; black lung disease; particle size distribution; Continuous personal dust monitor; Cascade Impactor; Electrical Low Pressure Impactor; Aerodynamic Particle Sizer; Aerodynamic Aerosol Classifier; Optical particle counter; DustTrak DRX

1. Introduction

Inhalation of respirable coal mine dust (RCMD) particles (with aerodynamic diameters \leq 4 micrometers [μ m]), and especially those containing quartz (crystalline silica), has been associated with coal workers' pneumoconiosis (CWP, sometimes referred to as "black lung") and silicosis diseases [1]. The extent, intensity, and constituents of RCMD exposure have been directly related to risks of human lung cellular damage and inflammation [2].

Implementation of the Mine Health and Safety Act of 1969 resulted in the reduction of RCMD mass and crystalline silica concentrations in U.S. mines $^{[3]}$. The National Institute of Occupational Safety and Health (NIOSH) $^{[3]}$ reported corresponding decreases in CWP occurrences for mid-central and south-central Appalachia underground coal miners between 1970 and 2000 (Figure 1). The 1969 regulation, along with improved mine ventilation, has resulted in reducing workplace disease $^{[4][5]}$. Since 2000, however, the prevalence and severity of RCMD-related lung diseases have increased $^{[6][7]}$, especially in mid-central Appalachia. New CWP and/or silicosis diagnoses are appearing in younger miners who should have benefitted from mine safety regulations $^{[8][9][10][11]}$. The 2014 Mine Safety and Health Administration's (MHSA) $^{[12]}$ respirable coal dust rule reduced permissible RCMD exposure from 2.0 to 1.5 mg/m³ over a full work shift. As a result, respirable dust sampling has gained importance for quantifying worker exposures and identifying RCMD sources. To improve measurement quality, MSHA $^{[12]}$ further requires the use of an approved continuous personal dust monitor (CPDM) from 2016 to complement the coal mine dust personal sampler unit (CMDPSU). As a comparison, the United States Environmental Protection Agency's (U.S. EPA) $^{[13]}$ national ambient air quality standards (NAAQS) for maximum 24 h PM2.5 (particles <2.5 µm aerodynamic diameter) exposure is 0.035 mg/m³.

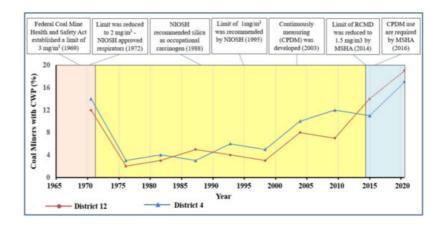


Figure 1. Coal workers' pneumoconiosis (CWP) prevalence in mid-central (District 4, Southern West Virginia) and south-central Appalachia (District 12) underground coal miners between 1970 and 2014. Data acquired from the Coal Workers' Health Surveillance Program (CWHSP) data query system [3][14], and includes all reported categories of CWP. CPDM: continuous personal dust monitor.

RCMD properties other than mass and crystalline silica, such as size, morphology, and chemical composition, also affect human health. Inhaled dust is deposited in different regions of the respiratory tract depending on particle sizes and shapes. The American Conference of Governmental Industrial Hygienists (ACGIH) established the following size fractions: (1) inhalable–particles capable of entering the nose and mouth; (2) thoracic–particles penetrating beyond the larynx; and (3) respirable–particles penetrating to the gas exchange region (alveolar) of the lung [15]. Size-selective sampling of these size fractions is defined by particle penetration efficiency curves, with 50% efficiencies at ~100, 10, and 4 μ m, respectively. In underground coal mines, dust particles \geq 1 μ m aerodynamic diameter from mechanical processes dominate the particle mass; however, ultrafine particles <0.1 μ m can dominate particle number concentrations in the presence of diesel engine exhaust [16]. Ultrafine particles present a health threat because of their potential to penetrate deep into the lung and pass across the air–blood barrier. Ultrafine particles present large surface areas that promote reactions with body fluids [17][18][19].

Mine safety regulations require RCMD mass to be collected by a size-selective cyclone inlet with a cut-off diameter (50% penetration efficiency) of ~4 μ m [20], which is an approximation of the inhalation properties of the human lung [21]. However, depending on particle size, RCMD mass can differ from the amount of dust that would deposit (i.e., dose) in the lung [22]. Size distribution measurements spanning the range from <0.1 μ m to >1 μ m particles are needed to assess potential health effects. Continuous size distributions enable the evaluation of metrics such as mass, surface area, and number concentration relationships to adverse health effects. Size distributions are also relevant to effective emission reduction measures, flammability, and explosive potential [23].

Most coal mine dust size distributions were collected over a decade ago [24][25], and may no longer be representative due to changes in underground coal mining conditions and practices. As coal seams become thinner, more rock strata (immediate roof and floor) are mined. Advances in longwall shearers have increased the volume of material handling, which can increase coal mine dust generation. It is important to understand how these changes affect particle size, shape, concentration, and composition. The National Academies of Sciences, Engineering, and Medicine (NASEM) [22] recommended a comprehensive characterization of RCMD chemical compositions and size distributions to identify additional causes of lung disease.

2. RCMD Mass Measurement Methods

Table 1 compares three commonly applied technologies for RCMD mass concentration measurement. Mine safety regulations require personal exposure monitoring in miners' breathing zones. In the U.S., RCMD mass concentrations are conventionally determined by sampling with a CMDPSU onto a filter followed by gravimetric analysis in a laboratory $^{[26]}$. As shown in Figure 2a, the CMDPSU is equipped with a belt-mounted constant-flow pump that draws air at 2 L per minute (L/min) through a 10-mm nylon Dorr–Oliver cyclone (or equivalent) and a pre-weighed filter. Under this flow rate, large particles with an aerodynamic diameter (dae) > 10 μ m are removed and collected in the cyclone hopper, which is cleaned between each use. Penetration efficiencies are ~50% for particles with dae ≈ 4 and 100% for dae < 2 μ m $^{[27]}$. As the cyclone sampling effectiveness curve $^{[28]}$ varies with flow rate, empirical conversion factors are applied to compensate for these changes. Due to the differences between the cyclone penetration- and respirable dust deposition-efficiency curves, conversion factors are also used to convert CMDPSU concentrations to other respirable dust conventions, such as the British Medical Research Council (BMRC) and the International Organization for Standardization (ISO) definitions $^{[29]}$.

Downstream of the cyclone, particles are collected onto a 37-mm-diameter polyvinyl chloride (PVC) filter with a pore size $\leq 5 \, \mu m$. The filters are sent to a laboratory for gravimetric and sometimes crystalline silica measurements.



Figure 2. Examples of RCMD mass measurement devices: (a) a coal mine dust personal sampler unit (CMDPSU) with a sampling pump, cyclone, and filter cassette; (b) ThermoScientific Personal Dust Monitor (PDM) 3700; and (c) TSI SidePak photometer personal aerosol monitor (SidePak has not been approved by the MSHA [Mine Safety and Health Administration] for use in underground coal mines).

Table 1. Comparison of RCMD mass concentration measurements.

Method	Description	Limitations and Challenges
Gravimetric sampler	Constant-flow sampling through a particle size- selective cyclone (e.g., Dorr–Oliver) onto a filter cartridge by a personal sampling pump The filter is submitted to gravimetric analysis and optionally for chemical analysis in the laboratory Reference method Relatively low cost	Ensuring that the cyclone assembly stays upright Labor intensive Low time resolution Data are not immediately available
Continuous personal dust monitor (CPDM)	A TEOM (tapered-element oscillating microbalance) obtains near real-time, gravimetric-equivalent measurement of RCMD mass concentrations Filter can be used for limited laboratory analysis Near real-time measurement (30-min average) Regulatory requirement Relatively independent of aerosol optical, physical, and chemical properties	High cost Size and weight are burdensome Regulatory requirement to report data to MSHA Potential evaporation losses
Photometer	Inferred mass concentration based on aerosol light scattering intensity Low cost Lightweight Fast response (~1 s)	Scattering-mass relationship varies with particle refractive index, shape, size distribution, density, and relative humidity Field calibration is needed

Filter sampling and gravimetric analysis [30] has been used as a reference method to demonstrate compliance with the RCMD exposure limit. However, it has several shortcomings. First, RCMD concentrations may take days or weeks to obtain, failing to provide critical information about the causes or prevention of overexposure. Second, the filter sample is collected over the entire shift and does not record temporal variations of RCMD exposures. Third, particle accumulation on cyclone walls, electrostatic charges, and cyclone orientation may affect the cyclone performance and introduce measurement bias [31][32][33]. The cyclone assembly must remain upright with the hopper facing downward, a stance difficult to maintain for the range of job activities. If the cyclone orientation is altered during measurement, oversized dust particles can deposit onto the filters, creating impurities that lead to inconclusive or inaccurate results.

Since February 2016, the use of a real-time CPDM for occupational exposure in high concentration areas and for miners with symptoms related to the development of pneumoconiosis has been required $^{[26]}$. The CPDM continuously measures RCMD mass concentrations and reports within-shift (30 min running average) and end-of-shift concentrations promptly upon the completion of the work shift $^{[26]}$. If RCMD concentrations exceed the permissible exposure limit, the mine operator is required to take immediate corrective actions. Presently, the only approved commercial CPDM is the personal dust monitor (PDM 3700, Thermo Fisher Scientific, Franklin, MA, USA; Figure 2b) $^{[34]}$. The PDM 3700 and its predecessor, PDM 3600, use a tapered-element oscillating microbalance (TEOM) to continuously measure the mass of collected particles $^{[35]}$. Particle-laden air is drawn through an inlet positioned in the miner's breathing zone at a flow rate of 2 L/min. The respirable dust is size-classified by a Higgens—Dewell type cyclone $^{[27][36]}$ and transported through a heated transfer line to the mass transducer worn at the miner's waist. Particles are collected on a Teflon-coated glass-fiber filter mounted on top of an oscillating hollow tapered element, for which the frequency decreases as particles deposit on the filter. The relationship between mass and tapered-element frequency changes is determined from calibration with known

masses [37]. The TEOM technology has been widely used in ambient particulate matter (PM) monitoring and is designated as a federal equivalent method by the U.S. EPA [38]. The PDM is a miniaturized version created for mining applications.

Laboratory and field measurements have evaluated PDM performance. Volkwein et al. [37][39] evaluated prototype PDMs in the laboratory using resuspended coal dust, finding that PDM mass concentrations were within ±25% of reference gravimetric measurements. Further field tests in ten coal mines found that the PDM had ~90% valid data availability for over 8000 h of underground use. Page et al. [40] conducted a linear regression of 129 pairs of PDM and CMDPSU measurements from 180 mechanized underground coal mining units and found the regression slope to be 1.05 (with zero intercept). Laboratory studies demonstrate that the PDM compares favorably with gravimetric mass concentrations for different aerosols, such as wood dust, aluminum oxide powder, flour dust, grain dust, diesel exhaust, welding fumes, Arizona road dust, and sodium chloride [41][42][43]. However, several studies indicate that transport losses and particle blow-off from the PDM filter may underestimate concentrations [41]. Loss of volatile material (as in diesel exhaust) due to the heating of the air inlet and tapered element to ensure stability may also result in negative biases for mass concentration.

The main PDM advantage is that it is comparable to gravimetric measurements and its response is independent of aerosol refractive index, size distribution, and density. The near real-time measurement provides miners with timely information to identify factors contributing to overexposures, allowing corrective actions to be taken immediately. In addition, the CPDM filters can be submitted to a laboratory for some chemical analyses [44][45]. However, its cost, size, and weight are drawbacks for routine use [22]. The high cost (~US\$17,000) limits the number of instruments used for purposes other than regulatory compliance. Currently, only a small fraction of miners wear CPDMs, causing concerns that many miners are insufficiently protected from dust exposure. The CPDM size and weight make the device burdensome to wear and the data are not easily observable by the miner. Furthermore, CPDM data must be reported to MSHA, which discourages mine operators from using CPDMs for noncompliance purposes, such as studying dust control effectiveness.

Different types of low-cost direct-reading dust monitors have been developed to supplement the regulatory required massbased CMDPSU and CPDM. Many of these monitors are photometers that use the principle of light scattering by an ensemble of particles to infer mass concentration [42][46][47][48]. As for the example shown in Figure 2c, a photometer draws particle laden air through a cyclone to achieve the desired size cut. The aerosol stream passes through a light beam, and the scattered light is measured at one or more scattering angles by photodetectors. Calibrated relationships are used to convert the scattered light intensity to particle mass concentration. Compared to CMDPSU and CPDM, photometers have the advantages of (1) low cost, (2) lighter weight, and (3) faster time response (as low as one second). Their main disadvantage is that the relationship between light scattering intensity and particle mass concentration depends on particle refractive index, shape, size distribution, density, and relative humidity [49]. Calibrations with collocated gravimetric measurements for different mining environments are needed, which are not always feasible. Although light scattering devices are not currently used for compliance with RCMD standards, they are still useful for dust source identification, emission control technology evaluation, and alerts for excessive exposure and the need to don personal protection equipment. Owing to their low cost, portability, low power requirements, and wireless communication potential, photometers can be installed in mining microenvironments to evaluate the temporal trends and spatial distributions of dust concentrations. Their lower cost and lighter weight allows them to be used by miners that are not required to wear a CPDM, allowing more miners' exposures to be monitored [22]. Photometers provide an opportunity to further develop wearable personal dust monitors with smaller size, lighter weight, and lower cost that can be provided to every miner for non-regulatory, and possible future regulatory, exposure assessments [50][51]. For application in underground coal mines, instruments should meet safety and other permissibility requirements for potentially explosive atmospheres, and they should be rugged enough to perform in a harsh mining environment without frequent maintenance.

3. RCMD Particle Size Characterization

Real-time airborne particle size distribution measurements have been reviewed by McMurry $^{[52]}$ for atmospheric aerosols and by Giechaskiel et al. $^{[53]}$ for engine emissions. Methods include microscopic imaging, aerodynamic sizing, optical sizing, and electrical mobility sizing. Pros and cons of each measurement method along with the detectable particle size ranges are summarized in <u>Table 2</u>. <u>Table S1</u> (supplemental) summarizes mining studies using these methods.

Table 2. A comparison of potential techniques that can be used for RCMD particle size characterization.

Technique	Advantages	Disadvantages
Optical Microscopy Size range > 1μm	Visual size and morphology evaluation	Time consuming; not suitable for submicron particles; potential observational bias and errors
SEM Size range: ~0.01–10 μm	Morphology and size analysis; elemental characteristics; wide particle size range	Laboratory measurement; needs sample pre- preparation; slow and costly; may not be representative as a small fraction of particles are analyzed
Cascade Impactor Size range: ~0.01–10 μm	Wide aerodynamic diameter range; size segregated mass concentration and chemical composition; can be used for personal sampling; mechanically rugged	Ex situ analysis; long sampling duration to collect sufficient mass; particle bounce may cause bias; non-uniform deposition
ELPI Size range: 0.006–10 μm	In situ real-time aerodynamic size distribution; wide size and concentration ranges	Particle bouncing; blow-off from substrates; overloading of substrates; low size resolution; charging efficiency uncertainty
APS Size range: 0.5–20 μm	In situ real-time aerodynamic size distributions; high size resolution; easy operation	Not suitable for particles <0.5 µm; density- dependent non-Stokesian correction; liquid particle deformation and losses; low concentration limit
AAC Size range: 0.025->5 μm	In situ aerodynamic size distributions; high size resolution; high transmission efficiency	Relatively slow scans (~2 min); fast rotating components; still under development/perfection
OPC Size range: ~0.3–10 μm	In situ real-time optical size distribution; compact and portable size; relatively low cost	Low concentration limit; dependence on particle shape and composition; non-monotonic dependence of light scattering on particle size
SMPS Size range: ~0.003–1 μm	In situ near real-time mobility size distribution; high size resolution and accuracy for submicron particles	Relatively slow scans; not suitable for >1 μm; limitation of using radioactive neutralizers
EEPS/FMPS/DMS Size range: 0.006–0.6 μm for EEPS and FMPS; 0.005– 2.5 μm for DMS	In situ real-time mobility size distribution; high time resolution; suitable for rapidly changing aerosols	Lower size resolution than SMPS; dependence of charging efficiency on particle morphology

3.1. Microscopic Imaging

Microscopic analysis can determine particle size and morphology. Image processing algorithms coupled with image libraries can classify particles by their shapes and textures, identify origins, and reveal potential inhalation and health consequences $^{[54]}$. A sufficient number of each particle type is required to represent exposure. Manual microscopic analysis is time consuming and requires user interpretation that may lead to observational biases and errors. Optical microscopy has been used to examine RCMD size distributions collected on filters or glass slides $^{[55][56]}$. However, submicron particles cannot be determined by optical methods due to the lower size limit of ~1 μ m.

Most modern RCMD applications use scanning electron microscopy (SEM) with a wider size range, from ~10 nm to tens of microns. Individual particle elemental compositions can be obtained when the SEM is equipped with an energy dispersive X-ray (EDX) detector. However, due to a lack of appropriate sampling substrates and skilled operators, RCMD has been only partially studied by this technique [16][57][58]. Computer-controlled SEM with EDX (CCSEM-EDX) reduces the personnel requirement and performs a frame-by-frame analysis for particle size, shape (e.g., aspect ratio), and elemental characteristics [16]. However, semi-volatile species evaporate under a vacuum, leading to biases for samples saturated with hydrocarbons (such as coal, organic materials, or swelling clays). "Low vacuum" and "environmental" SEMs are better suited for RCMD. Moreover, most SEMs use a fast-response solid state X-ray detector (Si(Li) detector), but render relatively low energy resolution and sensitivity for light elements (atomic number <12).

Efforts have been made to streamline SEM-EDX analysis for RCMD characterization [16][57][58][59]. Based on SEM imaging software, Sellaro et al. [58] used "line measurement" tools to find the long (L) and intermediate (I) particle dimensions while short (S) dimensions were estimated by assigning aspect ratios (longest divided by shortest dimension) for different minerals. The three dimensional parameters (i.e., L, I, and S) allow estimation of particle shape and volume. Based on their edge angles, particles can be classified as angular (a), transitional (t), or rounded (r) in shape, as shown in Figure 3 [60], which may be important for the particle's deposition and interaction with lung tissues. Particles are also classified into different categories based on elemental composition by EDX (e.g., quartz, carbonaceous, carbonate, etc.), each with an assumed aspect ratio and density. For reasonable analysis times (i.e., ~75–90 min per sample), Sellaro et al. [58]

recommended counting 100–200 particles with a magnification of $10,000\times$ to characterize ~0.5–8 µm dimensions. This method allows a large volume of samples to be analyzed cost-effectively. The example in <u>Figure 4</u> shows particle number concentration peaks at 0.5–1.0 µm followed by 1–2 µm with abundant aluminum-silica and mixed carbonaceous particles for the Roof Bolter sample.

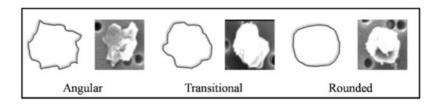


Figure 3. Angularity classification categories of SEM samples based on the qualitative analysis of the sharpness of particle edges [58].

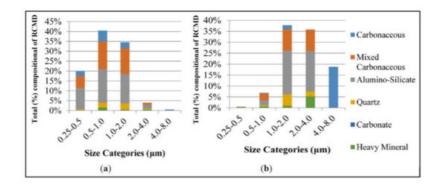


Figure 4. Particle size distribution by (a) number and (b) mass for the roof bolter sample (roof bolting machine); the relative number of particles in each compositional category is shown within each bar ^[58]. RCMD samples were collected from an underground coal mine in central Appalachia using a CMDPSU on 37 mm diameter polycarbonate filters. The center portion (9-mm diameter) of the filter was cut and attached to an SEM pin-stub for analysis.

Johann-Essex et al. ^[57] developed a CCSEM-EDX procedure using 1000× magnification to examine more than ~500 particles over a ~20 min sample analysis duration. However, the reduced image resolution only quantified particles ≥1µm. For the 209 samples collected from eight underground coal mines in three Appalachian regions, particles were classified into three size bins based on their cross-sectional diameters: 0.94–2.0, 2.0–3.0, and 3.0–9.0 µm, representing small, medium, and large RCMD, respectively. Particle sizes and aspect ratios varied among geological materials, mine operating conditions, and sampled microenvironments, with more of the smaller particles in the mid-central Appalachia mines and abundant elongated particles in the south-central Appalachia mines. Higher portions of fine (i.e., 0.94–2.0 µm) and elongated particles (i.e., aspect ratios between 1.5 and 3.0) were found at production faces and in return airways. Larger particles were found at feeders and intakes (e.g., surface resuspension) with dumping and vibrations, rather than active cutting. Particles with high carbonaceous (coal) content were larger and rounder than elongated alumino-silicate particles. High quartz content corresponded to smaller particles, while high carbonate content was found in rounder particles.

To characterize submicron particles (0.1–1 μ m), Sarver et al. [59] reanalyzed the Johann-Essex et al. [16] samples using a 20,000× magnification, finding that submicron particles dominated (>75%) the total particle number. In addition to diesel exhaust, cutting rock strata and rock dusting products were important fine particle sources. Sarver et al. [59] noted that the polycarbonate filter typically used in SEM analysis may have low collection efficiencies for submicron particles.

SEM-EDX limitations include the following: (1) the measurement is not in situ or real time; (2) particles are often collected on a polycarbonate filter and need to be transferred to an SEM stub and pretreated for analysis; (3) size-dependent particle collection efficiency and inhomogeneous deposition may lead to bias; (4) it is difficult to obtain optimal particle loadings; (5) only a fraction of collected particles are analyzed; (6) multiple magnifications are needed to analyze particles with a wide size range and it is difficult to obtain high-resolution images for particles <100 nm; and (7) the analysis is costly, time-consuming and may be subject to user interpretation.

3.2. Aerodynamic Particle Sizing

Aerodynamic particle sizing classifies particles based on aerodynamic diameter, which is defined as the diameter of a unit density sphere with the same settling velocity as the particle in question. The aerodynamic diameter is used to describe particle behavior in gravitational deposition, filtration, sampling, and penetration into the human respiratory system. Almost

all particle-related air quality standards and sampling conventions (e.g., $PM_{2.5}$, PM_{10} , and respirable dust) are defined based on aerodynamic rather than geometric diameters. Four types of aerodynamic particle sizing instruments relevant to RCMD measurement are as follows: (1) cascade impactors; (2) the electrical low pressure impactor (ELPI); (3) the aerodynamic particle sizer (APS); and (4) the aerodynamic aerosol classifier (AAC).

3.2.1. Cascade Impactors

Inertial cascade impactors cover size ranges from a few nanometers to tens of micrometers $\frac{[61]}{}$. Sequential impact stages accelerate the particle-laden flow through an array of jets positioned above flat substrates $\frac{[62]}{}$. Particles with aerodynamic diameters larger than the designed cut-off size deposit on the substrate, while smaller particles follow gas streamlines moving toward the next impaction stage. The impaction nozzles are progressively smaller with each stage, thereby accelerating the particle flow to higher velocities and collecting smaller particles. Substrates, such as aluminum foils, mylar sheets, and filters, can be placed on the impaction plate for offline laboratory analysis. A filter is placed at the last stage to collect remaining particles that are too small to impact. Each substrate is weighed before and after sampling to determine mass concentrations, thereby permitting mass-based size distributions to be determined using inversion techniques that incorporate the sampling effectiveness curves for each impactor stage $\frac{[63][64]}{}$. Subsequent chemical analyses of these substrates provide size-segregated chemical composition information.

Most RCMD size distributions were obtained using personal cascade impactors $\frac{[24][65][66][67]}{[24][55][66][67]}$. These small impactors with intrinsically safe sampling pumps have been worn by miners to estimate underground coal mine exposure. The Marple 290 series personal cascade impactor consists of up to eight impaction stages and a backup filter, with 50% cut-off diameters of 21.3, 14.8, 9.8, 6.0, 3.5, 1.55, 0.93, and 0.52 μ m at a flow rate of 2 L/min $\frac{[68]}{[68]}$. The Sioutas personal cascade impactor consists of four stages with diameters of 2.5, 1, 0.5, and 0.25 μ m at a flow rate of 9 L/min $\frac{[69]}{[69]}$. Chen et al. $\frac{[70]}{[69]}$ developed a 10 impaction stage personal impactor collecting 0.06–9.6 μ m particles at a flow rate of 2 L/min. Due to the low flow rate of personal cascade impactors (typically 2 L/min), a long sampling time is needed to collect sufficient mass for reliable gravimetric and chemical analyses. A 10 stage micro-orifice uniform deposit impactor (MOUDI) with 50% cut-point diameters of 10 to 0.056 μ m at 30 L/min, or the 13 stage MOUDI with cut-points of 10 to 0.010 μ m at flow rates of 10 or 30 L/min have been used to reduce sample durations, increase collected mass, and improve sizing resolution $\frac{[71]}{[72]}$. Eigure 5 shows mass-based size distributions measured in a diesel-powered coal mine having more submicron particles than in an entirely electric-powered coal mine, indicating the large contributions of diesel engine emissions to submicron particles $\frac{[72][73]}{[73]}$. An intrinsically safe pump is required to operate the MOUDI in underground coal mines.

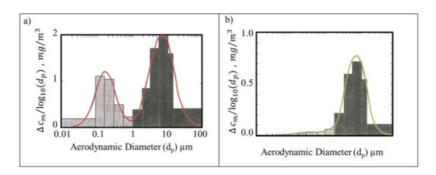


Figure 5. Mass-based size distribution measured by MOUDI at (a) a coal mine with diesel equipment, and (b) a coal mine with all electric equipment [72][73].

Particle bounce, wherein a larger deposited particle is re-entrained into the airflow for deposition on a subsequent stage, is a cause of uncertainty that shifts the distribution to smaller sizes compared to that in the atmosphere. Particles are not uniformly deposited across the impaction surface (except for the rotating MOUDI), with patterns reflecting the nozzle shapes and positions. As a result, filters cannot be sectioned for submission to different chemical analysis methods that assume a uniform deposition.

The quartz crystal microbalance coupled with a MOUDI (QCM-MOUDI) $^{[74]}$ determines real-time mass concentrations from the vibration frequency change of the quartz crystals. The QCM-MOUDI (model 140, TSI Inc., Shoreview, MN, USA) consists of a PM_{2.5} inlet and 6 QCM stages (2.44, 0.96, 0.51, 0.305, 0.156, 0.074, and 0.045 μ m). With a flow rate of 10 L/min, it records a mass-based size distribution every second, although longer integration periods are more accurate. Particle bounce is reduced by controlling the inlet flow relative humidity (RH) in the range of 40–65%. However, frequent cleaning is needed to prevent dust overloading and bounce in a harsh mining environment.

The ELPI (Dekati Ltd., Kangasala, Finland) measures particle size distribution as a function of aerodynamic diameter with high time resolution (as fast as every 0.1 s). Particles are sampled through a unipolar corona charger that imposes a well-defined charge distribution on particles independent of their initial charging state. The charged particles then pass through a low-pressure cascade impactor with electrically isolated collection stages. Particles are collected on impaction substrates depending on their aerodynamic diameters, and the electric charges carried by particles are measured by a multichannel electrometer $\frac{[75]}{[77][78]}$. Data inversion algorithms are used to convert raw current readings to particle number concentrations $\frac{[76][77][78]}{[78]}$. The ELPI consists of 14 impaction stages and 1 backup filter for particle sizes of 6 nm to 10 μ m. The main advantage of the ELPI for RCMD measurements is that it covers a wide size range with fast response times to distinguish rapidly changing nano- (e.g., DPM) and supermicron-particles. The aerodynamic diameter-based number distribution can be converted to mass distribution with less uncertainty due to variable particle density and shape. However, knowledge or assumptions of the effective density are required to reconcile differences between mobility diameter-dependent charging efficiency and aerodynamic diameter-dependent impaction separation $\frac{[79]}{[80]}$. ELPIs have been used in engine emission testing with good agreement between the ELPI and gravimetric mass concentrations $\frac{[80][81][82][83]}{[83]}$.

ELPIs have the same shortcomings as cascade impactors, such as particle bounce and blow-off from substrates and a relatively low size resolution. Particle charging efficiency depends on particle morphology, concentration, carrier gas composition, and relative humidity [84][85]. The aging and contamination of the charger can cause additional errors [86][87]. For RCMD measurement in underground coal mines, overloading of impaction stages, frequent cleaning, and intrinsic safety are the main limitations.

Bugarski et al. [88] used an ELPI to evaluate the effects of the longwall moving process in an underground trona mine (at an isolated zone test site) and found aerosol size distributions having two, three, or even four lognormal modes. As expected, diesel engines emitted submicron particles that dominated number concentrations, while supermicron dust particles dominated mass concentrations. Diesel-powered engines were found to be the primary source of both submicron aerosols and resuspended coarse dust. Although electrically powered vehicles did not directly generate tailpipe emissions, they also contributed to resuspended dust.

3.2.3. Aerodynamic Particle Sizer (APS)

When particles are rapidly accelerated through a nozzle, they attain different velocities depending on their inertia, which is a function of particle size and density $^{[89]}$. Smaller particles are accelerated faster due to their lower inertia. The APS measures the time-of-flight for particles passing through a path bounded by two laser beams of known separation downstream of an accelerating nozzle to infer velocities. The times-of-flight are then converted to aerodynamic diameters based on a calibration curve $^{[90]}$. The current APS (model 3321, TSI Inc., Shoreview, MN, USA) measures 52 size bins every second for aerodynamic diameters of 0.5–20 μ m with simultaneous optical sizing for 0.37–20 μ m diameter particles over a concentration range of ~0.001–1000 particles/cm 3 .

APS measurements have several limitations [91][92]. For larger particles, the acceleration velocity depends not only on aerodynamic diameter, but also on gas density, gas viscosity, particle density, and particle shape [93][94][95]. Corrections are possible when these properties are known; otherwise, the reported aerodynamic diameter may be biased [95]. The size of a particle with a density of 0.8 g/cm³ can be underestimated by as much as 5%, and a particle with a density of 2 g/cm³ can be overestimated by as much as 10%. Liquid droplets may deform during acceleration, leading to size underestimation. The degree of distortion depends on liquid viscosity and surface tension [96][97]. Liquid particles also have higher transport losses than solid particles [98]. The APS has a relatively low concentration limit (1000 particle/cm³) before coincidence errors (multiple particles passing through the laser beams) become significant. A dilutor can be used to reduce the particle number, but particle losses in the dilutor could lead to uncertainties for concentrations in the larger size channels.

The APS has been used in laboratory studies to measure the sampling effectiveness curves of aerodynamic classification devices such as cyclones $^{[27][33][36][99]}$ and impactors $^{[69][71]}$ that are used in mine applications. It has also been used for ambient aerosol size distributions, including locations close to mining areas $^{[100][101][102][103]}$. Concurrent mobility and aerodynamic size distribution measurements or using the APS to measure mobility classified particles allow for the estimation of particle effective densities and dynamic shape factors $^{[104][105][106]}$. Due to the low concentration range, the APS has had limited use for in-mine size distributions. Saarikoski et al. $^{[107]}$ used two scanning mobility particle sizers (SMPS) and an APS to measure size distributions in the range of 2.5 nm–10 μ m in an underground chrome mine. It was found that submicron particles from diesel engine exhaust and explosion combustion products yielded higher numbers and mass concentrations than mechanically generated coarse particles.

The AAC (Cambustion Ltd., Cambridge, UK) classifies particles based on a balance between centrifugal and drag forces $\frac{[108][109]}{[108]}$. The AAC consists of two concentric cylinders that spin at the same speed forming an annular classifying region. Aerosols enter the AAC near the wall of the inner cylinder, and traverse through clean sheath air towards the outer cylinder due to the centrifugal force. Particles with different inertias are separated into different trajectories: larger particles adhere to the outer cylinder wall; smaller particles exit with the excess sheath flow; and particles with the selected size exit through the monodisperse aerosol sampling port. By rotating the cylinders at different speeds, different particle sizes can be selected and measured by a condensation particle counter (CPC), generating number size distributions based on aerodynamic diameter. The AAC (1) covers a wide size range from 0.025 to >5 μ m; (2) does not require particle charging, resulting in transmission efficiencies 2–5 times higher than the SMPS that has an electrostatic operating principle; and (3) can be combined with mobility classification or mobility size distribution measurements to quantify particle effective densities, dynamic shape factors, and charge states $\frac{[110]}{100}$. Its drawbacks are that it takes several minutes to scan a size distribution, and the rotating components pose reliability challenges, particularly in harsh mining environments. The AAC is a relatively new instrument and its design is still being perfected for real-world size distribution measurements.

3.3. Optical Particle Sizing

Single particle optical particle counters (OPC) or spectrometers (OPS) measure particle sizes based on the amount of light scattered by individual particles, in contrast to photometers that measure total light scattering from an ensemble of particles $\frac{[52][111][112]}{}$. In an OPC, the light beam and particle stream are designed to reduce the probability of multiple particles being present in the sensing volume at the same time. The scattered light is converted to a proportional electrical pulse by a photodetector. The height or area of the pulse is used to infer particle diameter based on a predefined calibration curve, typically generated using spherical particles of known sizes and composition (e.g., polystyrene latex spheres). OPC designs differ in light sources (e.g., white light or wavelength-specific laser), scattered light collection angles (e.g., forward or side scattering), and photodetectors (e.g., photodiode or photomultiplier tube). Due to scattering by air molecules and electronic noise in the circuitry, most OPCs measure particle sizes in the range of ~0.3-10 µm. Advanced instruments can detect diameters as low as 0.05 µm [113]. Similar to photometers, the main advantages of OPCs are (1) fast time response (~1 s), (2) compact and portable size, and (3) relatively low cost. However, OPCs have low concentration limits (typically several thousand particles per cubic centimeter). Coincidence errors lead to inaccurate size and concentration measurements [114][115]. This problem can be partially overcome by combining single particle counting with photometry as implemented in the DustTrak DRX (TSI Inc., Shoreview, MN, USA), which measures sizesegregated concentrations up to 400 mg/m³ [49][116]. The light scattering signal depends not only on particle size, but also on particle refractive index and shape. Therefore, OPCs report "optical equivalent size" based on the particles used to establish the calibration curve, which may deviate from a particle's geometric, aerodynamic, or mobility size. The light scattering intensity vs. particle size curve is often not monotonic, especially for particles larger than 1 µm, leading to lower sizing resolutions and higher uncertainties [52].

OPC applications in mines have been limited, owing to the diverse refractive indices and non-spherical shapes of coal dust. Liu et al. [117] calibrated a near-forward scattering OPC by aerosolizing a small quantity of finely ground coal dust with a fluidized bed. A differential mobility analyzer (DMA) selected monodisperse particles over a size range from 0.4 to 2.4 µm. The experiments found that coal particles of the same mobility sizes generated lower OPC responses than transparent oil particles owing to the light-absorbing nature of the coal. The pulse height distributions from monodisperse coal particles were also broader than those for oil particles, likely due to their irregular shape. Without proper OPC calibration, coal dust sizes can be underestimated by up to fivefold [117][118]. However, OPC systematic sizing errors can be minimized by calibration with representative coal dusts. Barone et al. [119] applied ray tracing with diffraction on facets and T-matrix theories to adjust the responses of an OPC for submicron and micron size coal particles, respectively. This method accounted for the refractive index and non-spherical shape when computing coal dust diameters from light scattering theory. The size distributions measured by the OPC had reasonable agreement with APS, CCSEM, as well as cyclone-separated and sieve-segregated sizes. Marple and Rubow [120][121] calibrated an OPC to report aerodynamic sizes by measuring the sampling effectiveness curves of an impactor inlet and comparing to its known aerodynamic size penetration curve.

3.4. Electrical Mobility Particle Sizing

Electrical mobility is the most widely used technique to measure size distributions in the submicron size range [122]. There are two major designs: voltage scanning and non-scanning. The scanning mobility particle sizer (SMPS) consists of a bipolar charger, a differential mobility analyzer (DMA), and a condensation particle counter (CPC) [123]. Particles are first passed through a bipolar charger to obtain a well-defined stationary charge distribution [124][125]. The charger often uses a small quantity of radioactive material (e.g., krypton-85, polonium-210, or americalm-241) or soft X-rays to ionize molecules in the air, which subsequently attach to particles by diffusion charging [84][126][127][128][129]. Particles then enter

the DMA, where charged particles are separated into different trajectories by an electric field, depending on their electrical mobility [130]. At a given voltage, only particles of a given mobility size pass through the DMA to be counted by a CPC [131]. By varying the DMA voltage, particles with different sizes are selected and counted. The data inversion algorithm generates size distributions by taking into account the charge distribution, DMA transfer function, CPC counting efficiency, time constants, and particle transport losses [123][132][133]. Depending on DMA and CPC designs, the SMPS can quantify size distributions from several nanometers to several hundred nanometers every 1–2 min [134]. Recent advances in DMA, CPC, electrometers, and inversion algorithms include the following: (1) measuring particles down to 1 nm [135][136][137][138] [139][140][141][142], (2) measuring size distributions in less than one minute [133][143][144][145][146], and (3) more portable and rugged designs [147][148].

As the SMPS measures submicron particle size distributions, it is often used in parallel with an APS [106][112][149][150] or an OPC [151][152][153][154] to cover a wider size range. Skubacz et al. [155] used the SMPS and APS to measure 0.015–0.698 µm and 0.5–20 µm particle size distributions, respectively, in an underground coal mine. They observed high concentrations of ultrafine particles when electric-powered mining machines were in operation. Saarikoski et al. [156] combined an SMPS, OPC, and ELPI for particle size and a soot particle aerosol mass spectrometer (SP-AMS) for particle chemical composition. They found that engine exhaust emissions (dominated by organic matter and black carbon) accounted for 35–84% of submicron particle mass, and blasting (dominated by organic matter, sulfate, nitrate, ammonium, and black carbon) produced 7–60% of submicron particles' mass in an underground chrome mine.

In contrast to the SMPS, which scans voltage to obtain size distributions, a non-scanning mobility spectrometer uses multiple detectors to measure mobility-separated particle concentrations. Commercially available non-scanning instruments include the engine exhaust particle sizer spectrometer (EEPS; model 3090, TSI Inc., Shoreview, MN, USA), fast mobility particle sizer spectrometer (FMPS; model 3091, TSI Inc., Shoreview, MN, USA), and differential mobility spectrometer (DMS; Cambustion Ltd., Cambridge, UK). To increase detector signals, these instruments use unipolar chargers to charge the incoming particles, separate particles based on electrical mobility, and measure size-segregated particle concentrations using a series of electrometers [85][157][158][159][160][161][162]. These instruments can produce size distribution data as fast as every 0.1 s, and therefore are suitable for studying fast changing aerosols, such as in transient engine exhaust measurements. Their main disadvantages include (1) lower size resolution than the SMPS, (2) larger uncertainties in the unipolar charge distribution due to dependence on particle morphology [85][160], and (3) lower concentration sensitivity due to electrometer measurements.

Several studies have applied electrical mobility particle sizers in mining environments, particularly those related to engine exhaust. Bugarski and Hummer [163] used a FMPS to measure diesel-powered vehicle emissions in an underground mine to assess relative contributions of different types and categories of diesel engines to submicron aerosols and to assess the effectiveness of diesel emission control technologies. They found that replacing a U.S. EPA pre-tier engine with a Tier 3 engine resulted in 41% lower particle number concentrations. A retrofitted disposable filter element reduced particle number emissions by 77–92%, while a retrofit sintered metal filter reduced particle emissions by 93%. Bugarski et al. [164] also used an FMPS and an ELPI to measure size distirbutions emitted by a diesel-powered personnel carrier vehicle and by a manual metal arc welder in an underground mine. The FMPS size distributions for both diesel exhast and welding fumes had modes at ~10 nm and ~70 nm, with welding aerosols having an additional mode at ~140 nm. The ELPI data demonstrated that neither diesel exhaust nor welding generated micron-sized particles.

3.5. Evaluation for Size Distribution Measurements in Mines

The ELPI has several advantages for RCMD size distribution measurements. First, it measures a wide aerodynamic diameter range from 6 nm to 10 μm, covering both diesel exhaust and mechanically generated dust particles. Second, it has a high time resolution (0.1 s), allowing it to measure size distributions of fast changing aerosols. Third, because the ELPI measures aerodynamic diameter, the conversion from number distribution to mass distribution has less influence from particle properties than SMPS or OPCs, making the integrated mass concentrations more comparable to the regulatory required gravimetric mass concentrations. A recently developed ELPI algorithm reports high resolution (up to 500 size channels) inverted size distributions [78]. The full size distribution allows for calculations of size-segregated particle surface and mass concentrations, permitting evaluation of the effects of these alternative metrics on RCMD health effects. Different substrates can be used in the ELPI to collect particles for microscopic and chemical analyses. In addition to ELPI, cascade impactors and SEM-EDX analysis can complement RCMD characterization. Cascade impactors do not only derive mass distributions, they also allow analysis of particle chemical compositions in different size ranges. SEM-EDX analysis provides additional information about particle morphology and particle-level chemical composition.

4. Summary

Although adverse effects of RCMD on workers' health have been recognized for decades and several regulations and research efforts have been focused on this issue, there is an increasing prevalence and severity of coal mine dust-related lung diseases in some regions. This review assesses measurement technologies that characterize coal mine dust mass concentrations and size distributions for mining areas. Comparisons of different techniques are summarized with examples where these methods have been applied (with a focus on U.S. coal mines). Some of the advanced instrument presented in this paper are not intrinsically safe (e.g., ELPI, APS, and AAC) and caution should be exerted when using them in explosive environments. Outlines for per-forming comprehensive characterization of RCMD mass and size distribution are recommended. This review indicates that many coal mine dust size distributions are decades old and may not represent modern mining technologies (e.g., increased equipment size and power and mining thinner coal seams). It is apparent that RCMD and silica exposures need to be supplemented with more detailed chemical knowledge of potentially toxic species. Future studies are essential to provide insights into the causes for recent increases in coal miner lung diseases.

References

- 1. Castranova, V.; Vallyathan, V. Silicosis and coal workers' pneumoconiosis. Environ. Health Perspect. 2000, 108, 675–684.
- 2. Halldin, C.N.; Wolfe, A.L.; Laney, A.S. Comparative respiratory morbidity of former and current US coal miners. Am. J. Public Health 2015, 105, 2576–2577.
- 3. Centers for Disease Control and Prevention. Coal Workers' Health Surveillance Program—Data Query System; NIOSH: Morgantown, WV, USA, 2016.
- 4. Antao, V.C.; Petsonk, E.L.; Attfield, M.D. Advanced Cases of Coal Workers' Pneumoconiosis—Two Counties, Virginia, 2006. JAMA 2006, 296, 2085.
- 5. Fitzhugh, P.; Margach, E. Department of Health and Human Services Coal mine dust exposures and associated health outcomes- A review. In Coal Mine Dust: Health Reviews and Key Studies (with accompanying CD); Nova Science Publishers Inc: Hauppauge, NY, USA, 2013; pp. 25–67. ISBN 9781624170973.
- 6. Laney, A.S.; Petsonk, E.L.; Attfield, M.D. Pneumoconiosis among underground bituminous coal miners in the United States: Is silicosis becoming more frequent? Occup. Environ. Med. 2010, 67, 652–656.
- 7. Blackley, D.J.; Reynolds, L.E.; Short, C.; Carson, R.; Storey, E.; Halldin, C.N.; Laney, A.S. Progressive massive fibrosis in coal miners from 3 clinics in Virginia. JAMA 2018, 319, 500–501.
- 8. Blackley, D.J.; Crum, J.B.; Halldin, C.N.; Storey, E.; Laney, A.S. Resurgence of progressive massive fibrosis in coal miners Eastern Kentucky, 2016. Morb. Mortal. Wkly. Rep. 2016, 65, 1385–1389.
- 9. Laney, A.S.; Petsonk, E.L.; Hale, J.M.; Wolfe, A.L.; Attfield, M.D. Potential determinants of coal workers' pneumoconiosis, advanced pneumoconiosis, and progressive massive fibrosis among underground coal miners in the United States, 2005-2009. Am. J. Public Health 2012, 102, 2005–2009.
- 10. Suarthana, E.; Laney, A.S.; Storey, E.; Hale, J.M.; Attfield, M.D. Coal workers' pneumoconiosis in the United States: Regional differences 40 years after implementation of the 1969 Federal Coal Mine Health and Safety Act. Occup. Environ. Med. 2011, 68, 908–913.
- 11. Department of Health and Human Services Coal Mine Dust Exposures and Associated Health Outcomes: A Review of Information Published Since 1995; National Institute for Occupational Safety and Health (NIOSH): Morgantown, WV, USA, 2011; Volume 64.
- 12. MSHA Lowering Miners' Exposure to Respirable Coal Mine Dust, Including Continuous Personal Dust Monitors. Fed. Regist. 2014, 79, 24814–24994.
- 13. U.S. EPA National ambient air quality standards for particulate matter: Proposed rule. Fed. Regist. 2012, 77, 38890–39055.
- 14. Madl, A.K.; Donovan, E.P.; Gaffney, S.H.; McKinley, M.A.; Moody, E.C.; Henshaw, J.L.; Paustenbach, D.J. State-of-the-science review of the occupational health hazards of crystalline silica in abrasive blasting operations and related requirements for respiratory protection. J. Toxicol. Environ. Health. B. Crit. Rev. 2008, 11, 548–608.
- 15. Vincent, J. Particle size-selective sampling for particulate air contaminants; American Conference of Governmental Industrial Hygienists (ACGIH): Cincinnati, OH, USA, 1999; ISBN 978-1882417308.

- 16. Johann-Essex, V.; Keles, C.; Sarver, E. A Computer-Controlled SEM-EDX Routine for Characterizing Respirable Coal Mine Dust. Minerals 2017, 7, 15.
- 17. Kelly, F.J.; Fussell, J.C. Size, source and chemical composition as determinants of toxicity attributable to ambient particulate matter. Atmos. Environ. 2012, 60, 504–526.
- 18. Oberdörster, G.; Oberdörster, E.; Oberdörster, J. Nanotoxicology: An emerging discipline evolving from studies of ultrafine particles. Environ. Health Perspect. 2005, 113, 823–839.
- 19. Riediker, M.; Zink, D.; Kreyling, W.; Oberdörster, G.; Elder, A.; Graham, U.; Lynch, I.; Duschl, A.; Ichihara, G.; Ichihara, S.; et al. Particle toxicology and health where are we? Part. Fibre Toxicol. 2019, 16, 19.
- 20. Gautam, M.; Sreenath, A. Performance of a respirable multi-inlet cyclone sampler. J. Aerosol Sci. 1997, 28, 1265–1281.
- 21. Vincent, J.H. Health-related aerosol measurement: A review of existing sampling criteria and proposals for new ones. J. Environ. Monit. 2005, 7, 1037–1053.
- 22. National Academies of Sciences, Engineering, and Medicine (NASEM). Monitoring and Sampling Approaches to Assess Underground Coal Mine Dust Exposures; National Academies Press: Washington, DC, USA, 2018.
- 23. Li, Q.; Wang, K.; Zheng, Y.; Ruan, M.; Mei, X.; Lin, B. Experimental research of particle size and size dispersity on the explosibility characteristics of coal dust. Powder Technol. 2016, 292, 290–297.
- 24. Birch, M.E.; Noll, J.D. Submicrometer elemental carbon as a selective measure of diesel particulate matter in coal mines. J. Environ. Monit. 2004, 6, 799–806.
- 25. Burkhart, J.E.; McCawley, M.A.; Wheeler, R.W. Particle Size Distributions in Underground Coal Mines. Am. Ind. Hyg. Assoc. J. 1987, 48, 122–126.
- 26. Mine Safety and Health Administration (MSHA). 30 CFR Part 74: Coal Mine Dust Sampling Devices, Federal Register; U.S. Department of Labor: Washington, DC, USA, 2016.
- 27. Bartley, D.L.; Chen, C.C.; Song, R.; Fischbach, T.J. Respirable Aerosol Sampler Performance Testing. Am. Ind. Hyg. Assoc. J. 1994, 55, 1036–1046.
- 28. Watson, J.G.; Chow, J.C.; Shah, J.J.; Pace, T.G. The effect of sampling inlets on the PM10 and PM15 to TSP concentration ratios. J. Air Pollut. Control Assoc. 1983, 33, 114–119.
- 29. Page, S.J.; Volkwein, J.C. A revised conversion factor relating respirable dust concentrations measured by 10 mm Dorr-Oliver nylon cyclones operated at 1.7 and 2.0 L min-1. J. Environ. Monit. 2009, 11, 684.
- 30. Watson, J.G.; Tropp, R.J.; Kohl, S.D.; Wang, X.L.; Chow, J.C. Filter processing and gravimetric analysis for suspended particulate matter samples. Aerosol Sci. Eng. 2017, 1, 193–205.
- 31. Blachman, M.W.; Lippmann, M. Performance Characteristics of the Multicyclone Aerosol Sampler. Am. Ind. Hyg. Assoc. J. 1974, 35, 311–326.
- 32. Kar, K.; Gautam, M. Orientation Bias of the Isolated 10-mm Nylon Cyclone at Low Stream Velocity. Am. Ind. Hyg. Assoc. J. 1995, 56, 1090–1098.
- 33. Tsai, C.-J.; Shiau, H.-G.; Lin, K.-C.; Shih, T.-S. Effect of deposited particles and particle charge on the penetration of small sampling cyclones. J. Aerosol Sci. 1999, 30, 313–323.
- 34. Thermo Fisher Scientific. Personal Dust Monitor (Model PDM3700) Instruction Manual; Thermo Fisher Scientific: Franklin, MA, USA, 2016.
- 35. Patashnick, H.; Rupprecht, E.G. Continuous PM-10 Measurements Using the Tapered Element Oscillating Microbalance. J. Air Waste Manag.. Assoc. 1991, 41, 1079–1083.
- 36. Maynard, A.D.; Kenny, L.C. Performance assessment of three personal cyclone models, using an Aerodynamic Particle Sizer. J. Aerosol Sci. 1995, 26, 671–684.
- 37. Volkwein, J.C.; Vinson, R.P.; McWilliams, L.J.; Tuchman, D.P.; Mischler, S.E. Mining Publication: Performance of a New Personal Respirable Dust Monitor for Mine Use, Report of Investigations 9663; National Institute for Occupational Safety and Health (NIOSH): Pittsburgh, PA, USA, 2004.
- 38. Chow, J.C.; Fairley, D.; Watson, J.G.; de Mandel, R.; Fujita, E.M.; Lowenthal, D.H.; Lu, Z.; Frazier, C.A.; Long, G.; Cordova, J. Source apportionment of wintertime PM10 at San Jose, CA. J. Environ. Eng 1995, 21, 378–387.
- 39. Volkwein, J.C.; Vinson, R.P.; Page, S.J.; McWilliams, L.J.; Joy, G.J.; Mischler, S.E.; Tuchman, D.P. Report of investigations (National Institute for Occupational Safety and Health. In Laboratory and Field Performance of a Continuously Measuring Personal Respirable Dust Monitor; Department of Healthy and Human Services: Cincinnati, OH, USA, 2006.

- 40. Page, S.J.; Volkwein, J.C.; Vinson, R.P.; Joy, G.J.; Mischler, S.E.; Tuchman, D.P.; McWilliams, L.J. Equivalency of a personal dust monitor to the current United States coal mine respirable dust sampler. J. Environ. Monit. 2008, 10, 96–101
- 41. Halterman, A.; Sousan, S.; Peters, T.M. Comparison of Respirable Mass Concentrations Measured by a Personal Dust Monitor and a Personal DataRAM to Gravimetric Measurements. Ann. Work Expo. Heal. 2017, 62, 62–71.
- 42. Thorpe, A.; Walsh, P.T. Direct-Reading Inhalable Dust Monitoring—An Assessment of Current Measurement Methods. Ann. Occup. Hyg. 2013, 57, 824–841.
- 43. Noll, J.; Volkwein, J.; Janisko, S.; Patts, L. Portable instruments for measuring tailpipe diesel particulate in underground mines. Min. Eng. 2013, 65, 42–49.
- 44. Bulpitt, S.; Price, M. The composition of PM10 as collected by a conventional TEOM, a modified TEOM and a Partisol gravimetric monitor at a kerbside site in the North East of England. Environ. Monit. Assess. 2006, 121, 479–489.
- 45. Nosratabadi, A.R.; Graff, P.; Karlsson, H.; Ljungman, A.; Leanderson, P. Use of TEOM monitors for continuous long-term sampling of ambient particles for analysis of constituents and biological effects. Air Qual. Atmos. Heal. 2019, 12, 161–171
- 46. Patts, J.R.; Tuchman, D.P.; Rubinstein, E.N.; Cauda, E.G.; Cecala, A.B. Performance Comparison of Real-Time Light Scattering Dust Monitors Across Dust Types and Humidity Levels. Min. Met. Explor. 2019, 36, 741–749.
- 47. Pui, D.Y.H. Direct-reading instrumentation for workplace aerosol measurements. A review. Analyst 1996, 121, 1215.
- 48. Thorpe, A. Assessment of Personal Direct-Reading Dust Monitors for the Measurement of Airborne Inhalable Dust. Ann. Occup. Hyg. 2007, 51, 97–112.
- 49. Wang, X.L.; Chancellor, G.; Evenstad, J.; Farnsworth, J.E.; Hase, A.; Olson, G.M.; Sreenath, A.; Agarwal, J.K. A Novel Optical Instrument for Estimating Size Segregated Aerosol Mass Concentration in Real Time. Aerosol Sci. Technol. 2009, 43, 939–950.
- 50. Hagler, G.S.W.; Solomon, P.A.; Hunt, S.W. New Technology for Low-Cost, Real-Time Air Monitoring; Air and Waste Management Association's Magazine for Environmental Managers; Air & Waste Management Association: Pittsburgh, PA, USA, 2014.
- 51. Morawska, L.; Thai, P.K.; Liu, X.; Asumadu-Sakyi, A.; Ayoko, G.; Bartonova, A.; Bedini, A.; Chai, F.; Christensen, B.; Dunbabin, M.; et al. Applications of low-cost sensing technologies for air quality monitoring and exposure assessment: How far have they gone? Environ. Int. 2018, 116, 286–299.
- 52. McMurry, P.H. A Review of Atmospheric Aerosol Measurement. Atmos. Environ. 2000, 34, 1959–1999.
- 53. Giechaskiel, B.; Maricq, M.; Ntziachristos, L.; Dardiotis, C.; Wang, X.; Axmann, H.; Bergmann, A.; Schindler, W. Review of motor vehicle particulate emissions sampling and measurement: From smoke and filter mass to particle number. J. Aerosol Sci. 2014, 67, 48–86.
- 54. Shamir, L.; Delaney, J.D.; Orlov, N.; Eckley, D.M.; Goldberg, I.G. Pattern recognition software and techniques for biological image analysis. PLoS Comput. Biol. 2010, 6, e1000974.
- 55. Cartwright, J. Relationships between Mass and Number Concentrations of Respirable Airborne Dust in British Coal Mines: Part 1—Mass-Number Relationships for Coal-Dust Clouds. Ann. Occup. Hyg. 1965, 8, 193–206.
- 56. Stein, F.; Corn, M. Shape factors of narrow size range samples of respirable coal mine dust. Powder Technol. 1975, 13, 133–141.
- 57. Johann-essex, V.; Keles, C.; Rezaee, M.; Scaggs-witte, M.; Sarver, E. International Journal of Coal Geology Respirable coal mine dust characteristics in samples collected in central and northern Appalachia. Int. J. Coal Geol. 2017, 182, 85–93.
- 58. Sellaro, R.; Sarver, E.; Baxter, D. A Standard Characterization Methodology for Respirable Coal Mine Dust Using SEM-EDX. Resources 2015, 939–957.
- 59. Sarver, E.; Keles, C.; Rezaee, M. International Journal of Coal Geology Beyond conventional metrics: Comprehensive characterization of respirable coal mine dust. Int. J. Coal Geol. 2019, 207, 84–95.
- 60. Powers, M.C. A New Roundness Scale for Sedimentary Particles. Sepm J. Sediment. Res. 1953, 23, 117-119.
- 61. Arffman, A.; Kuuluvainen, H.; Harra, J.; Vuorinen, O.; Juuti, P.; Yli-Ojanperä, J.; Mäkelä, J.M.; Keskinen, J. The critical velocity of rebound determined for sub-micron silver particles with a variable nozzle area impactor. J. Aerosol Sci. 2015, 86, 32–43.
- 62. Marple, V.A. History of Impactors—The First 110 Years. Aerosol Sci. Technol. 2004, 38, 247–292.

- 63. Ramachandran, G.; Johnson, E.W.; Vincent, J.H. Inversion techniques for personal cascade impactor data. J. Aerosol Sci. 1996, 27, 1083–1097.
- 64. Winklmayr, W.; Wang, H.-C.; John, W. Adaptation of the Twomey Algorithm to the Inversion of Cascade Impactor Data. Aerosol Sci. Technol. 1990, 13, 322–331.
- 65. Potts, J.D.; McCawley, M.A.; Jankowski, R.A. Thoracic Dust Exposures on Longwall and Continuous Mining Sections. Appl. Occup. Environ. Hyg. 1990, 5, 440–447.
- 66. Rubow, K.L.; Cantrell, B.K.; Marple, V.A. Measurement of coal dust and diesel exhaust aerosols in underground mines. In Proceedings of the 7th International Pneumoconioses Conference, Pittsburgh, PA, USA, 23–26 August 1988; pp. 23–26.
- 67. Seixas, N.S.; Hewett, P.; Robins, T.G.; Haney, R. Variability of Particle Size-Specific Fractions of Personal Coal Mine Dust Exposures. Am. Ind. Hyg. Assoc. J. 1995, 56, 243–250.
- 68. Rubow, K.L.; Marple, V.A.; Olin, J.; McCawley, M.A. A Personal Cascade Impactor Design, Evaluation and Calibration. AIHAJ 1987, 48, 532–538.
- 69. Misra, C.; Singh, M.; Shen, S.; Sioutas, C.; Hall, P.M. Development and evaluation of a personal cascade impactor sampler (PCIS). J. Aerosol Sci. 2002, 33, 1027–1047.
- 70. Chen, M.; Romay, F.J.; Marple, V.A. Design and evaluation of a low flow personal cascade impactor. Aerosol Sci. Technol. 2018, 52, 192–197.
- 71. Marple, V.A.; Rubow, K.L.; Behm, S.M. A Microorifice Uniform Deposit Impactor (MOUDI): Description, Calibration, and Use. Aerosol Sci. Technol. 1991, 14, 434–446.
- 72. Cantrell, B.K. Source apportionment analysis applied to mine dust aerosols: Coal dust and diesel emissions aerosol measurement. In Proceedings of the 3rd Mine Ventilation Symposium, University Park, PA, USA, 12–14 October 1987; Society of Mining Engineering: Littleton, CO, USA, 1987; pp. 495–501.
- 73. Cantrell, B.K.; Volkwein, J.C. Mine Aerosol Measurement. In Aerosol Measurement: Principles, Techniques, and Applications; Baron, P.A., Willeke, K., Eds.; John Wiley & Sons, Inc.: New York, NY, USA, 2001; pp. 801–820.
- 74. Chen, M.; Romay, F.J.; Li, L.; Naqwi, A.; Marple, V.A. A novel quartz crystal cascade impactor for real-time aerosol mass distribution measurement measurement. Aerosol Sci. Technol. 2016, 50, 971–983.
- 75. Keskinen, J.; Pietarinen, K.; Lehtimäki, M. Electrical low pressure impactor. J. Aerosol Sci. 1992, 23, 353–360.
- 76. Lemmetty, M.; Keskinen, J.; Marjamäki, M. The ELPI Response and Data Reduction II: Properties of Kernels and Data Inversion. Aerosol Sci. Technol. 2005, 39, 583–595.
- 77. Marjamäki, M.; Lemmetty, M.; Keskinen, J. ELPI Response and Data Reduction I: Response Functions. Aerosol Sci. Technol. 2005, 39, 575–582.
- 78. Saari, S.; Arffman, A.; Harra, J.; Rönkkö, T.; Keskinen, J. Performance evaluation of the HR-ELPI + inversion. Aerosol Sci. Technol. 2018, 52, 1037–1047.
- 79. Moisio, M. Real time size distribution measurement of combustion aerosols. Ph.D. Thesis, Tampere University of Technology, Tampere, Finland, 1999.
- 80. Lehmann, U.; Niemelä, V.; Mohr, M. New Method for Time-Resolved Diesel Engine Exhaust Particle Mass Measurement. Environ. Sci. Technol. 2004, 38, 5704–5711.
- 81. Maricq, M.M.; Szente, J.; Loos, M.; Vogt, R. Motor Vehicle PM Emissions Measurement at LEV III Levels. SAE Int. J. Engines 2011, 4, 597–609.
- 82. Maricq, M.M. Monitoring Motor Vehicle PM Emissions: An Evaluation of Three Portable Low-Cost Aerosol Instruments. Aerosol Sci. Technol. 2013, 47, 564–573.
- 83. Maricq, M.M.; Szente, J.J.; Harwell, A.L.; Loos, M.J. How well can aerosol instruments measure particulate mass and solid particle number in engine exhaust? Aerosol Sci. Technol. 2016, 50, 605–614.
- 84. Jiang, J.; Kim, C.; Wang, X.; Stolzenburg, M.R.; Kaufman, S.L.; Qi, C.; Sem, G.J.; Sakurai, H.; Hama, N.; McMurry, P.H. Aerosol Charge Fractions Downstream of Six Bipolar Chargers: Effects of Ion Source, Source Activity, and Flowrate. Aerosol Sci. Technol. 2014, 48, 1207–1216.
- 85. Wang, X.L.; Grose, M.A.; Caldow, R.; Osmondson, B.L.; Swanson, J.J.; Chow, J.C.; Watson, J.G.; Kittelson, D.B.; Li, Y.; Xue, J.; et al. Improvement of Engine Exhaust Particle Sizer (EEPS) size distribution measurement: II. Engine exhaust aerosols. J. Aerosol Sci. 2016, 92, 83–94.
- 86. Järvinen, A.; Aitomaa, M.; Rostedt, A.; Keskinen, J.; Yli-Ojanperä, J. Calibration of the new electrical low pressure impactor (ELPI+). J. Aerosol Sci. 2014, 69, 150–159.

- 87. Marjamäki, M.; Keskinen, J.; Chen, D.-R.; Pui, D.Y.H. Performance evaluation of the electrical low-pressure impactor (ELPI). J. Aerosol Sci. 2000, 31, 249–261.
- 88. Bugarski, A.D.; Hummer, J.A.; Vanderslice, S.; Shahan, M.R. Characterization of Aerosols in an Underground Mine during a Longwall Move. Min. Met. Explor. 2020, 37, 1065–1078.
- 89. Wilson, J.C.; Liu, B.Y.H. Aerodynamic particle size measurement by laser-doppler velocimetry. J. Aerosol Sci. 1980, 11, 139–150.
- 90. Agarwal, J.K.; Remiarz, R.J.; Quant, F.R.; Sem, G.J. Real-time aerodynamic particle size analyzer. J. Aerosol Sci 1982, 13, 222–223.
- 91. Baron, P.A. Calibration and Use of the Aerodynamic Particle Sizer (APS 3300). Aerosol Sci. Technol. 1986, 5, 55-67.
- 92. Baron, P.A.; Mazumder, M.K.; Cheng, Y.-S.; Peters, T.M. Real-Time Techniques for Aerodynamic Size Measurement. Aerosol Meas. 2011, 36, 313–338.
- 93. Ananth, G.; Wilson, J.C. Theoretical Analysis of the Performance of the TSI Aerodynamic Particle Sizer The Effect of Density on Response. Aerosol Sci. Technol. 1988, 9, 189–199.
- 94. Rader, D.J.; Brockmann, J.E.; Ceman, D.L.; Lucero, D.A. A Method To Employ the Aerodynamic Particle Sizer Factory Calibration Under Different Operating Conditions. Aerosol Sci. Technol. 1990, 13, 514–521.
- 95. Wang, H.C.; John, W. Particle Density Correction for the Aerodynamic Particle Sizer. Aerosol Sci. Technol. 1987, 6, 191–198.
- 96. Baron, P.; Deye, G.J.; Martinez, A.B.; Jones, E.N.; Bennett, J.S. Size Shifts in Measurements of Droplets with the Aerodynamic Particle Sizer and the Aerosizer. Aerosol Sci. Technol. 2008, 42, 201–209.
- 97. Bartley, D.L.; Martinez, A.B.; Baron, P.A.; Secker, D.R.; Hirst, E. Droplet Distortion in Accelerating Flow. J. Aerosol Sci. 2000, 31, 1447–1460.
- 98. Volckens, J.; Peters, T.M. Counting and particle transmission efficiency of the aerodynamic particle sizer. J. Aerosol Sci. 2005, 36, 1400–1408.
- 99. Chen, C.C.; Huang, S.H. Shift of Aerosol Penetration in Respirable Cyclone Samplers. Am. Ind. Hyg. Assoc. J. 1999, 60, 720–729.
- 100. Kurth, L.M.; McCawley, M.; Hendryx, M.; Lusk, S. Atmospheric particulate matter size distribution and concentration in West Virginia coal mining and non-mining areas. J. Expo. Sci. Environ. Epidemiol. 2014, 24, 405–411.
- 101. Peters, T.M. Use of the Aerodynamic Particle Sizer to Measure Ambient PM10–2.5: The Coarse Fraction of PM10. J. Air Waste Manag. Assoc. 2006, 56, 411–416.
- 102. Reid, J.S.; Jonsson, H.H.; Maring, H.B.; Smirnov, A.; Savoie, D.L.; Cliff, S.S.; Reid, E.A.; Livingston, J.M.; Meier, M.M.; Dubovik, O.; et al. Comparison of size and morphological measurements of coarse mode dust particles from Africa. J. Geophys. Res. 2003, 108, 8593.
- 103. Sioutas, C. Evaluation of the Measurement Performance of the Scanning Mobility Particle Sizer and Aerodynamic Particle Sizer. Aerosol Sci. Technol. 1999, 30, 84–92.
- 104. DeCarlo, P.; Slowik, J.; Worsnop, D.; Davidovits, P.; Jimenez, J. Particle Morphology and Density Characterization by Combined Mobility and Aerodynamic Diameter Measurements. Part 1: Theory. Aerosol Sci. Technol. 2004, 38, 1185– 1205.
- 105. Park, K.; Dutcher, D.; Emery, M.; Pagels, J.; Sakurai, H.; Scheckman, J.; Qian, S.; Stolzenburg, M.R.; Wang, X.; Yang, J.; et al. Tandem Measurements of Aerosol Properties—A Review of Mobility Techniques with Extensions. Aerosol Sci. Technol. 2008, 42, 801–816.
- 106. Stanier, C.O.; Khlystov, A.Y.; Pandis, S.N. Ambient aerosol size distributions and number concentrations measured during the Pittsburgh Air Quality Study (PAQS). Atmos. Environ. 2004, 38, 3275–3284.
- 107. Saarikoski, S.; Teinilä, K.; Timonen, H.; Aurela, M.; Laaksovirta, T.; Reyes, F.; Vásques, Y.; Oyola, P.; Artaxo, P.; Pennanen, A.S.; et al. Particulate matter characteristics, dynamics, and sources in an underground mine. Aerosol Sci. Technol. 2018, 52, 114–122.
- 108. Johnson, T.J.; Irwin, M.; Symonds, J.P.R.; Olfert, J.S.; Boies, A.M. Measuring aerosol size distributions with the aerodynamic aerosol classifier. Aerosol Sci. Technol. 2018, 52, 655–665.
- 109. Tavakoli, F.; Olfert, J.S. An Instrument for the Classification of Aerosols by Particle Relaxation Time: Theoretical Models of the Aerodynamic Aerosol Classifier. Aerosol Sci. Technol. 2013, 47, 916–926.
- 110. Tavakoli, F.; Olfert, J.S. Determination of particle mass, effective density, mass—mobility exponent, and dynamic shape factor using an aerodynamic aerosol classifier and a differential mobility analyzer in tandem. J. Aerosol Sci. 2014, 75,

- 111. International Organization for Standardization (ISO). ISO 21501-1:2009 Determination of Particle Size Distribution—Single Particle Light Interaction Methods—Part 1: Light Scattering Aerosol Spectrometer; ISO: Geneva, Switzerland, 2009
- 112. Sorensen, C.M.; Gebhart, J.; O'Hern, T.J.; Rader, D.J. Optical Measurement Techniques: Fundamentals and Applications. In Aerosol Measurement; John Wiley & Sons, Inc.: Hoboken, NJ, USA, 2011; pp. 269–312. ISBN 9781118001684.
- 113. Pinnick, R.G.; Pendleton, J.D.; Videen, G. Response Characteristics of the Particle Measuring Systems Active Scattering Aerosol Spectrometer Probes. Aerosol Sci. Technol. 2000, 33, 334–352.
- 114. Whitby, K.T.; Liu, B.Y.H. Generation of countable pulses by high concentrations of subcountable sized particles in the sensing volume of optical counters. J. Colloid Interface Sci. 1967, 25, 537–546.
- 115. Whitby, K.T.; Willeke, K. Single particle optical counters: Principles and field use. In Aerosol Measurement; Lundgren, D.A., Lippmann, M., Harris, F.S., Clark, W.E., Marlow, W.H., Durham, M.D., Eds.; University Presses of Florida: Gainesville, FL, USA, 1979; pp. 145–182.
- 116. Wang, X.; Zhou, H.; Arnott, W.P.; Meyer, M.E.; Taylor, S.; Firouzkouhi, H.; Moosmüller, H.; Chow, J.C.; Watson, J.G. Evaluation of gas and particle sensors for detecting spacecraft-relevant fire emissions. Fire Saf. J. 2020, 113, 102977.
- 117. Liu, B.Y.H.; Marpple, V.A.; Whitby, K.T.; Barsic, N.J. Size Distribution Measurement of Airborne Coal Dust by Optical Particle Counters. Am. Ind. Hyg. Assoc. J. 1974, 35, 443–451.
- 118. Whitby, K.T.; Vomela, R.A. Response of single particle optical counters to nonideal particles. Environ. Sci. Technol. 1967, 1, 801–814.
- 119. Barone, T.L.; Hesse, E.; Seaman, C.E.; Baran, A.J.; Beck, T.W.; Harris, M.L.; Jaques, P.A.; Lee, T.; Mischler, S.E. Calibration of the cloud and aerosol spectrometer for coal dust composition and morphology. Adv. Powder Technol. 2019, 30, 1805–1814.
- 120. Marple, V.A.; Rubow, K.L. Aerodynamic particle size calibration of optical particle counters. J. Aerosol Sci. 1976, 7, 425–433.
- 121. Marple, V.A.; Rubow, K.L. A portable optical particle counter system for measuring dust aerosols. Am. Ind. Hyg. Assoc. J. 1978, 39, 210–218.
- 122. Flagan, R.C. History of Electrical Aerosol Measurements. Aerosol Sci. Technol. 1998, 28, 301–380.
- 123. Wang, S.C.; Flagan, R.C. Scanning Electrical Mobility Spectrometer. Aerosol Sci. Technol. 1990, 13, 230-240.
- 124. Fuchs, N.A. On the stationary charge distribution on aerosol particles in a bipolar ionic atmosphere. Geofis. Pura E Appl. 1963, 56, 185–193.
- 125. Wiedensohler, A. An approximation of the bipolar charge distribution for particles in the submicron size range. J. Aerosol Sci. 1988, 19, 387–389.
- 126. Kallinger, P.; Steiner, G.; Szymanski, W.W. Characterization of four different bipolar charging devices for nanoparticle charge conditioning. J. Nanoparticle Res. 2012, 14, 944.
- 127. Lee, H.M.; Soo Kim, C.; Shimada, M.; Okuyama, K. Bipolar diffusion charging for aerosol nanoparticle measurement using a soft X-ray charger. J. Aerosol Sci. 2005, 36, 813–829.
- 128. Liu, B.Y.H.; Pui, D.Y.H. Equilibrium bipolar charge distribution of aerosols. J. Colloid Interface Sci. 1974, 49, 305-312.
- 129. Liu, B.Y.H.; Pui, D.Y.H. Electrical neutralization of aerosols. J. Aerosol Sci. 1974, 5, 465–472.
- 130. Knutson, E.O.; Whitby, K.T. Aerosol classification by electric mobility: Apparatus, theory, and applications. J. Aerosol Sci. 1975, 6, 443–451.
- 131. McMurry, P.H. The History of Condensation Nucleus Counters. Aerosol Sci. Technol. 2000, 33, 297-322.
- 132. Mai, H.; Flagan, R.C. Scanning DMA Data Analysis I. Classification Transfer Function. Aerosol Sci. Technol. 2018, 52, 1382–1399.
- 133. Mai, H.; Kong, W.; Seinfeld, J.H.; Flagan, R.C. Scanning DMA data analysis II. Integrated DMA-CPC instrument response and data inversion. Aerosol Sci. Technol. 2018, 52, 1400–1414.
- 134. Watson, J.G.; Chow, J.C.; Sodeman, D.A.; Lowenthal, D.H.; Chang, M.-C.O.; Park, K.; Wang, X.L. Comparison of four scanning mobility particle sizers at the Fresno Supersite. Particuology 2011, 9, 204–209.
- 135. Brunelli, N.A.; Flagan, R.C.; Giapis, K.P. Radial Differential Mobility Analyzer for One Nanometer Particle Classification. Aerosol Sci. Technol. 2009, 43, 53–59.

- 136. Cai, R.; Chen, D.-R.; Hao, J.; Jiang, J. A miniature cylindrical differential mobility analyzer for sub-3 nm particle sizing. J. Aerosol Sci. 2017, 106, 111–119.
- 137. Franchin, A.; Downard, A.; Kangasluoma, J.; Nieminen, T.; Lehtipalo, K.; Steiner, G.; Manninen, H.E.; Petäjä, T.; Flagan, R.C.; Kulmala, M. A new high-transmission inlet for the Caltech nano-RDMA for size distribution measurements of sub-3nm ions at ambient concentrations. Atmos. Meas. Tech. 2016, 9, 2709–2720.
- 138. lida, K.; Stolzenburg, M.R.; McMurry, P.H. Effect of Working Fluid on Sub-2 nm Particle Detection with a Laminar Flow Ultrafine Condensation Particle Counter. Aerosol Sci. Technol. 2008, 43, 81–96.
- 139. Jiang, J.; Chen, M.; Kuang, C.; Attoui, M.; McMurry, P.H. Electrical Mobility Spectrometer Using a Diethylene Glycol Condensation Particle Counter for Measurement of Aerosol Size Distributions Down to 1 nm. Aerosol Sci. Technol. 2011, 45, 510–521.
- 140. Kangasluoma, J.; Cai, R.; Jiang, J.; Deng, C.; Stolzenburg, D.; Ahonen, L.R.; Chan, T.; Fu, Y.; Kim, C.; Laurila, T.M.; et al. Overview of measurements and current instrumentation for 1–10 nm aerosol particle number size distributions. J. Aerosol Sci. 2020, 148, 105584.
- 141. Kuang, C.; Chen, M.; McMurry, P.H.; Wang, J. Modification of Laminar Flow Ultrafine Condensation Particle Counters for the Enhanced Detection of 1 nm Condensation Nuclei. Aerosol Sci. Technol. 2012, 46, 309–315.
- 142. Stolzenburg, M.R.; Scheckman, J.H.T.; Attoui, M.; Han, H.-S.; McMurry, P.H. Characterization of the TSI model 3086 differential mobility analyzer for classifying aerosols down to 1 nm. Aerosol Sci. Technol. 2018, 52, 748–756.
- 143. Enroth, J.; Kangasluoma, J.; Korhonen, F.; Hering, S.; Picard, D.; Lewis, G.; Attoui, M.; Petäjä, T. On the time response determination of condensation particle counters. Aerosol Sci. Technol. 2018, 52, 778–787.
- 144. Fernandez de la Mora, J.; Perez-Lorenzo, L.J.; Arranz, G.; Amo-Gonzalez, M.; Burtscher, H. Fast high-resolution nanoDMA measurements with a 25 ms response time electrometer. Aerosol Sci. Technol. 2017, 51, 724–734.
- 145. Shah, S.D.; Cocker, D.R. A Fast Scanning Mobility Particle Spectrometer for Monitoring Transient Particle Size Distributions. Aerosol Sci. Technol. 2005, 39, 519–526.
- 146. Tröstl, J.; Tritscher, T.; Bischof, O.F.; Horn, H.-G.; Krinke, T.; Baltensperger, U.; Gysel, M. Fast and precise measurement in the sub-20nm size range using a Scanning Mobility Particle Sizer. J. Aerosol Sci. 2015, 87, 75–87.
- 147. Amanatidis, S.; Kim, C.; Spielman, S.R.; Lewis, G.S.; Hering, S.V.; Flagan, R.C. The Spider DMA: A miniature radial differential mobility analyzer. Aerosol Sci. Technol. 2020, 54, 175–189.
- 148. Hering, S.V.; Lewis, G.S.; Spielman, S.R.; Eiguren-Fernandez, A. A MAGIC concept for self-sustained, water-based, ultrafine particle counting. Aerosol Sci. Technol. 2019, 53, 63–72.
- 149. Khlystov, A.; Stanier, C.; Pandis, S.N. An Algorithm for Combining Electrical Mobility and Aerodynamic Size Distributions Data when Measuring Ambient Aerosol Special Issue of Aerosol Science and Technologyon Findings from the Fine Particulate Matter Supersites Program. Aerosol Sci. Technol. 2004, 38, 229–238.
- 150. Liu, J.; Jiang, J.; Zhang, Q.; Deng, J.; Hao, J. A spectrometer for measuring particle size distributions in the range of 3 nm to 10 μm. Front. Environ. Sci. Eng. 2016, 10, 63–72.
- 151. Hand, J.L.; Kreidenweis, S.M. A New Method for Retrieving Particle Refractive Index and Effective Density from Aerosol Size Distribution Data. Aerosol Sci. Technol. 2002, 36, 1012–1026.
- 152. Wang, X.L.; Watson, J.G.; Chow, J.C.; Gronstal, S.; Kohl, S.D. An efficient multipollutant system for measuring real-world emissions from stationary and mobile sources. Aerosol Air Qual. Res. 2012, 12, 145–160.
- 153. Watson, J.G.; Chow, J.C.; Lowenthal, D.H.; Magliano, K.L. Estimating aerosol light scattering at the Fresno Supersite. Atmos. Environ. 2008, 42, 1186–1196.
- 154. Woo, K.-S. Measurement of Atmospheric Aerosols: Size Distribution of Nanoparticles, Estimation of Distribution Moments and Control of Relative Humidity; University of Minnesota: Minneapolis, MN, USA, 2002.
- 155. Skubacz, K.; Wojtecki, Ł.; Urban, P. Aerosol concentration and particle size distributions in underground excavations of a hard coal mine. Int. J. Occup. Saf. Erg. 2017, 23, 318–327.
- 156. Saarikoski, S.; Salo, L.; Bloss, M.; Alanen, J.; Teinilä, K.; Reyes, F.; Vázquezuez, Y.; Keskinen, J.; Oyola, P.; Rönkkö, T.; et al. Sources and Characteristics of Particulate Matter at Five Locations in an Underground Mine. Aerosol Air Qual. Res. 2019, 9, 2613–2624.
- 157. Biskos, G.; Reavell, K.; Collings, N. Description and Theoretical Analysis of a Differential Mobility Spectrometer. Aerosol Sci. Technol. 2005, 39, 527–541.
- 158. Johnson, T.; Caldow, R.; Pöcher, A.; Mirme, A.; Kittelson, D. A New Electrical Mobility Particle Sizer Spectrometer for Engine Exhaust Particle Measurements; SAE International: Warrendale, PA, USA, 2004.

- 159. Symonds, J.P.R.; Reavell, K.S.J.; Olfert, J.S.; Campbell, B.W.; Swift, S.J. Diesel soot mass calculation in real-time with a differential mobility spectrometer. J. Aerosol Sci. 2007, 38, 52–68.
- 160. Wang, X.L.; Grose, M.A.; Avenido, A.; Stolzenburg, M.R.; Caldow, R.; Osmondson, B.L.; Chow, J.C.; Watson, J.G. Improvement of Engine Exhaust Particle Sizer (EEPS) size distribution measurement: I. Algorithm and Applications to Compact Aerosols. J. Aerosol Sci. 2016, 92, 95–108.
- 161. Nakamura, K.; Oshikiri, T.; Ueno, K.; Wang, Y.; Kamata, Y.; Kotake, Y.; Misawa, H. Properties of Plasmon-Induced Photoelectric Conversion on a TiO2 /NiO p–n Junction with Au Nanoparticles. J. Phys. Chem. Lett. 2016, 7, 1004–1009.
- 162. Reavell, K.; Hands, T.; Collings, N. A Fast Response Particulate Spectrometer for Combustion Aerosols. SAE Technical Paper2002-01-2714. Sae Trans. 2002, 111, 1338–1344.
- 163. Bugarski, A.D.; Hummer, J.A. Contribution of various types and categories of diesel-powered vehicles to aerosols in an underground mine. J. Occup. Environ. Hyg. 2020, 17, 121–134.
- 164. Bugarski, A.D.; Barone, T.L.; Hummer, J.A. Diesel and welding aerosols in an underground mine. Int. J. Min. Sci. Technol. 2020.

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